

Modeling considerations for research on Ocean Alkalinity Enhancement (OAE)

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Abstract

The deliberate increase of ocean alkalinity (referred to as Ocean Alkalinity Enhancement or OAE) has been proposed as a method for removing CO₂ from the atmosphere. Before OAE can be implemented safely, efficiently, and at scale several research questions have to be addressed including: 1) which alkaline feedstocks are best suited and in what doses can they be added safely, 2) how can net carbon uptake be measured and verified, and 3) what are the potential ecosystem impacts. These research questions cannot be addressed by direct observation alone but will require skillful and fit-for-purpose models. This [article](#) provides an overview of the most relevant modeling tools, including turbulence-, regional- and global-scale biogeochemical models, and techniques including approaches for model validation, data assimilation, and uncertainty estimation. Typical biogeochemical model assumptions and their limitations are discussed in the context of OAE research, which leads to an identification of further development needs to make models more applicable to OAE research questions. A description of typical steps in model validation is followed by proposed minimum criteria for what constitutes a model that is fit for its intended purpose. After providing an overview of approaches for sound integration of models and observations via data assimilation, the application of Observing System Simulation Experiments (OSSEs) for observing system design is described within the context of OAE research. Criteria for model validation and intercomparison studies are presented. The article concludes with a summary of recommendations and potential pitfalls to be avoided.

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41 1 Introduction

42 Ocean Alkalinity Enhancement (OAE) refers to the deliberate increase of ocean alkalinity, which
43 can be realized either by removing acidic substances from or adding alkaline substances to
44 seawater. OAE is receiving increasing attention as a method for removing CO₂ from the
45 atmosphere; such methods are referred to as marine Carbon Dioxide Removal (mCDR)
46 technologies (Renforth and Henderson, 2017). Natural analogues to OAE exist (Shubas et al.
47 2023). An increase in the alkalinity of seawater leads to a repartitioning of its dissolved
48 carbonate species with a shift toward bicarbonate and carbonate ions (Zeebe and Wolf-Gladrow
49 2001, Renforth and Henderson 2017), leading to a reduction in the aqueous CO₂ concentration
50 and thus the partial pressure of CO₂ ($p\text{CO}_2$; Schulz et al. 2023). Since exchange of CO₂ between
51 the ocean and atmosphere occurs when the surface ocean $p\text{CO}_2$ is out of equilibrium with that of
52 the atmosphere, a lowering of the ocean's $p\text{CO}_2$ will lead to a net ingassing of atmospheric CO₂
53 (i.e., an increase in CO₂ uptake by the ocean or a decrease in outgassing due to OAE). This
54 would increase the oceanic and decrease the atmospheric inventories of inorganic carbon, in
55 other words, it would result in mCDR. In contrast to other mCDR technologies, OAE does not
56 exacerbate ocean acidification (Ilyina et al. 2013). In fact, an increase in ocean alkalinity
57 counteracts acidification, and while subsequent net uptake of atmospheric CO₂ largely restores
58 pH to its pre-perturbation value, there is potential for OAE deployment to mitigate acidification
59 impacts near injection sites (Mongin et al. 2021).

60 Several important research questions should be addressed before implementing OAE as an
61 mCDR technology at scale. These include: 1) which alkaline substances are best suited and in
62 what doses can they be added reliably while avoiding precipitation of calcium carbonate (which
63 would decrease alkalinity and could result in runaway precipitation events), 2) how can
64 changes in alkalinity and net carbon uptake be measured, verified, and reported (referred to as
65 MRV; see Ho et al. 2023) to enable meaningful carbon crediting, and 3) what are the potential
66 ecosystem impacts and how can harm to ecosystems be avoided or minimized while
67 maximizing potential benefits. These research questions cannot be addressed by direct
68 observation alone, but will require an integration of observations and numerical ocean models
69 across a range of scales. Skillful and fit-for-purpose models will be essential for addressing
70 many OAE research questions including the MRV challenge, assessment of environmental
71 impacts, and interpretation of natural analogs.

72 Ocean models are useful for a broad range of purposes, from idealized models for basic
73 hypothesis testing of fundamental principles to realistic models for more applied uses (see
74 primer on ocean biogeochemical models by Fennel et al. 2022). In the context of OAE research,
75 this full range of models is applicable. For example, idealized models of particle-fluid
76 interaction can inform us about dissolution and precipitation kinetics at the scale of particles,
77 realistic local-scale models can inform us about nearfield processes in the turbulent
78 environment around injection sites, and larger-scale regional or global ocean models can be
79 used to support observational design for field experiments, to demonstrate possible verification
80 frameworks, and to address questions about global-scale feedbacks on ocean biogeochemistry.
81 A common objective of all these modeling approaches is to realistically simulate the spatio-

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temporal evolution of the seawater carbon chemistry, including alkalinity and dissolved CO₂, and attribute that evolution to physical, chemical, and biological processes. Models that are suitable for this purpose will provide spatial and temporal context for properties that can be observed (but at much sparser temporal and spatial coverage than a model can provide) as well as estimates of properties and fluxes that cannot be directly observed but may be inferred because of known mechanistic relationships or patterns of correlation. Applications of realistic models rely on them being skillful and accurate, requiring that they include parameterizations of the relevant processes, and that they are constrained by observations that contain sufficient meaningful information (what is sufficient depends on the application and research question). Methods for constraining models by observations through statistically optimal combination of both are available. Application of such methods is referred to as data assimilation and provides the most accurate estimates of biogeochemical properties and fluxes (see Fennel et al. 2022 for fundamentals and code examples).

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Model applications for OAE research include the following four general types:

- Hindcasts are model applications where a defined time period in the past was simulated. They can be unconstrained—in the sense that no observations are fed into the model except for initial, boundary, and forcing conditions—or constrained, where observations inform the model state via data assimilation. The latter are also referred to as optimal hindcasts or reanalyses.
- Nowcasts/forecasts are similar to constrained hindcasts but with the simulations carried out up to the present (referred to as nowcasts) or into the future (referred to as forecasts). The latter require assumptions about future forcing and boundary conditions, e.g., from other forecasts, climatology, or assuming persistence.
- Scenarios are unconstrained hindcasts or forecasts where one or more aspects of the model is systematically perturbed to assess the effect of the perturbation, for example, in paired simulations with and without OAE, one would be the realistic case and the other, a scenario (also referred to as counterfactual in this case). These can be used to explore even very unlikely situations, which is often required in comprehensive uncertainty and risk assessment.
- Observing System Simulation Experiments (OSSEs) for observing system design use unconstrained and/or constrained hindcasts to evaluate the benefits of different sampling designs and optimize deployment of observational assets for a defined objective, including tradeoffs between different types of observation platforms.

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Successful implementation of models to support OAE research and MRV is challenging because of the general sparseness of relevant biogeochemical observations, and the limited lab, mesocosm, and field trial data available to date for model parameterization. Further, models are built at a process level and integrated to reveal behavior at the emergent scale. As such, models comprise a collective hypothesis of the ocean's physical, biogeochemical, and ecosystem function—but it is important to recognize that model formulations of key processes related to

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OAE remain uncertain. It may well turn out that parameterizations of the carbonate system, of plankton diversity and trophic interactions, small scale turbulence, submesoscale subduction and restratification processes, and air-sea gas exchange in the current generation of models require improvement to robustly treat OAE-related questions.

The intended scope of this [article](#) is to provide an overview of the most relevant modeling tools for OAE research with high-level background information, illustrative examples, and references to more in-depth methodological descriptions and further examples. We aim to provide simple criteria and guidance for researchers on the current state-of-the-art of biogeochemical modeling relevant to OAE research, keeping in mind short-term research goals in support of pilot deployments of OAE and long-term goals such as credible MRV in an ocean affected by large-scale deployment of OAE and possibly other CDR technologies.

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2 Modeling approaches

This section provides a brief review of modeling tools available for OAE research with references to more in-depth methodological descriptions and examples, as well as a discussion of which approaches are most applicable to simulating essential processes in different circumstances. The presentation is structured using two complementary organizing principles, the spatial and temporal scales of the problem in Section 2.1 and the biogeochemical and ecological complexity represented by different modeling approaches in Section 2.2. [Section 2](#), concludes with a summary of suggested future model development efforts in Section 2.3.

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2.1. Modeling approaches across scales

In the nearfield, close to the site of an alkalinity increase, an accurate characterization of the spatio-temporal evolution of alkalized waters requires direct representation or parameterization of fluid and particle physics and seawater carbonate chemistry at scales ranging from micrometers to hundreds of meters, spanning turbulent to submesoscale processes (Section 2.1.1). In the farfield, covering scales from 10s of meters to 100s of kilometers, where the effect of an alkalinity increase depends less on the details of how the alkalinity was added, or acidity removed, and is instead dominated by ambient environmental processes, local to regional scale models are useful for simulating the impact of alkalinity increases, for verifying the intended perturbations in air-sea exchange of CO₂ and in carbonate system variables, and potentially for simulating ecosystem impacts (Section 2.1.2). Lastly, investigation of the effects of the global ocean's overturning circulation, impacts on atmospheric CO₂ levels, and of Earth system feedbacks resulting from deployment of OAE and other CDR technology at scale requires global modeling approaches (Section 2.1.3).

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2.1.1. Particle scale to nearfield/turbulence scale (μm to km scales)

Small-scale modeling approaches cover the range from μm-size particles to the turbulent- and submeso-scales in the nearfield of alkalinity additions. Simulating processes on these scales allows one to address questions about how turbulent mixing dilutes and disperses alkalized water and how it affects the settling, aggregation, disaggregation, precipitation, and dissolution

179 of suspended particles. Nearfield modeling has an important role to play in guiding the design
180 of deployment strategies that mitigate environmental impacts and meet future permitting
181 requirements, and to support monitoring. During the initial dispersion and dilution phase of an
182 alkalinity increase in the nearfield, the direct impacts on carbonate system variables are
183 greatest, with waters exhibiting the largest elevations in pH and the highest potential for the
184 formation of secondary precipitates. For particulate alkalinity feedstocks, turbulence close to the
185 deployment site affects dissolution and settling rates, increasing dissolution and either
186 accelerating or diminishing the settling of sedimentary particles compared to the Stokes settling
187 speed (Fornari et al. 2016).

188 Distinct approaches to modeling at these scales involve different levels of parametrization and
189 computational expense, with the relative utility of each approach being dependent on the
190 scientific questions at hand. At the smallest scales, Direct Numerical Simulations (DNS) are the
191 most computationally expensive and specialized class of fluid modeling, as they resolve flows
192 down to the scales at which flow variances dissipate—typically centimeters or smaller in the
193 ocean. Consequently, computational constraints imply that they cannot be run over domains
194 larger than a few meters. DNS are thus integrated over idealized physical domains (i.e., they
195 lack realistic bathymetry) and are suited to investigating fundamental physical processes. For
196 example, multiphase DNS simulations have been used to model the interaction of turbulence
197 with gas bubbles (Farsoiya et al. 2023) and particles (Fornari et al. 2016). Results from such
198 studies provide an important testbed that can be used to develop parameterizations required in
199 lower resolution models.

200 A well-established approach to modeling the fluid flow at scales up to about 10 km uses Large
201 Eddy Simulations (LES), a class of model that directly solves the unsteady Navier-Stokes
202 equations down to the largest turbulent scales on a high-resolution grid. Such models
203 parameterize turbulence using a subgrid-scale model (e.g., Smagorinsky 1963). An advantage of
204 these models is their ability to simulate both an alkalized plume and the environmental
205 turbulence into which the plume emerges. Once alkalized waters enter the surface boundary
206 layer, LES models have an established history of simulating turbulence and mixing that is
207 directly relevant to OAE research (e.g., Mensa et al. 2015, Taylor et al. 2020). An example of an
208 LES simulation of near-surface turbulence dispersing surface-deployed alkalinity downwards is
209 illustrated in Figure 1, where a physical model (Ramadhan et al. 2020) has been coupled to a
210 carbonate solver (Lewis et al. 1998). To date, LES models have rarely been coupled to
211 biogeochemical models due to the computational expenses involved, though their inclusion
212 may be increasingly feasible (Smith et al. 2018, Whitt et al. 2019). As LES simulate flow physics
213 at scales ranging from 10-10,000 m, they do not explicitly resolve the microscales of fluid motion
214 and chemical reactions at particle scales. Nevertheless, the parameterizations of such processes
215 can be included; for example, Liang et al. (2011) used models of bubble concentration and
216 dissolved gas concentration in an LES to examine the influence of bubbles on air-sea gas
217 exchange.

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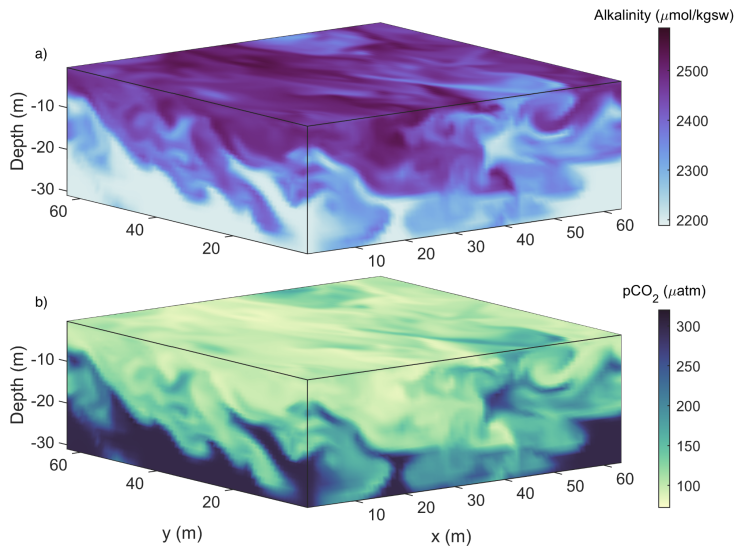


Figure 1: LES of near surface turbulence coupled to a carbonate system solver. Alkalinity is added at a rate of $4 \mu\text{mol kgsw}^{-1} \text{m}^{-2} \text{s}^{-1}$ for 20 minutes to the top grid cell at the start of the simulation. Turbulence, generated by surface wind stress and cooling, sets the rate at which it mixes downwards (a) along with associated waters of lowered $p\text{CO}_2$ (b). Turbulent plumes and eddies lead to inhomogeneities in water properties at scales of tens of meters.

For alkalized plumes associated with outfalls from, for example, wastewater treatment plants, integral models (that assume plume properties such that the governing equations are simplified) have been developed to examine the initial dilution close to jets and buoyant plumes up to kilometer scales (Jirka et al. 1996). These models are highly configurable, enabling specific diffuser configurations as well as the potential to incorporate sediment laden plumes with particle settling (Bleninger & Jirka 2004). Results are commonly accepted for engineering purposes, defining mixing zones, and providing a fast “first look” at diffusion and mixing near an outfall site. However, these models rely on assumptions about the underlying physics of fluid flow (e.g., axisymmetric plumes and simplified entrainment rates) that may not be accurate under general oceanic conditions, and results will not include all effects of irregular bathymetry, finite domain size or arbitrarily non-uniform ambient conditions. Nevertheless, their simplicity makes them very useful. For example, by combining several simple process models for plume dilution, particle dissolution, and carbon chemistry, Caserini et al. (2021) have simulated the initial dilution of slaked lime $\text{Ca}(\text{OH})_2$ particles and alkalinity in a plume behind a moving vessel.

Other methods for modeling at this scale include Reynolds Averaged Navier Stokes (RANS) and Unsteady RANS (URANS), wherein fluctuations against a slowly varying or time mean

background are parametrized, often using constant (large) eddy diffusivities and viscosities. These approaches are often inaccurate at these scales, resulting in simulations that are too diffusive or lacking processes that are of leading order importance to mixing (Golshan et al. 2017, Chang & Scotti 2004).

There are multiple, potentially interacting sources of uncertainty to consider when evaluating the uncertainty of the applications described above. Perhaps best understood but still problematic is the uncertainty that arises from the computational intractability of simulating all the relevant scales in the μm to km range at once, necessitating the different modeling approaches for different scales, with parameterizations to account for unresolved scales and scale interactions. The dissolved carbonate chemistry of seawater is relatively well parameterized (Zeebe and Wolf-Gladrow 2001), but some modest uncertainties arise from approximations required for computational tractability (Smith et al. 2018). The least understood but potentially dominant source of uncertainty pertains to the representation of the microscale biological, chemical, and physical dynamics of particles, which is an active area of experimental and observational investigation (Subhas et al. 2022, Fuhr et al. 2022, Hartmann et al. 2023).

While the explicit multiphase modeling of the particles themselves is computationally costly, an approach wherein the parametrized evolution of inertia-less Lagrangian particles are simulated may provide a fruitful middle ground, providing a mechanism to realistically determine the alkalinity release field associated with the advection, mixing, sinking and dissolution of reactive mineral particles. These questions about particles apply to those released in OAE deployments, as well as particles that precipitate from seawater in part due to OAE deployments, and finally the role of ambient biotic and abiotic particles where OAE is deployed.

2.1.2. Local to regional scales (m to km)

Local to regional scale models that range in horizontal resolution from tens of meters to hundreds of kilometers are useful for simulating the impact of alkalinity injections beyond the immediate local area, where conditions do not depend on the details of how the alkalinity was added and instead are determined by regional-scale currents and other process, including the potential for biogenic feedbacks. These models are particularly useful to support OAE field experiments, including planning and observational design, and analysis, integration and synthesis of observations, and to facilitate interpretation of observations from natural analogs. Furthermore, local and regional scale models will likely prove to be indispensable for quantification of OAE effects in research settings, for guiding assessments of its environmental impacts, and for MRV during the potential implementation of OAE. A skillful model can simulate when and where changes in carbonate chemistry and the ensuing anomalies in air-sea CO_2 exchange occur and provide an estimate of the spatio-temporal extent of the biogeochemical properties affected by OAE.

Regional models have distinct advantages over global models in their ability to resolve the spatial scales on which OAE would be applied both experimentally and operationally, and their documented skill in representing coastal and continental shelf processes more accurately (Mongin et al. 2016, Laurent et al. 2021). Examples of regional model applications in the context

285 of OAE include the recent studies by Mongin et al. (2021) and Wang et al. (2023). Mongin et al.
286 (2021) used a coupled physical-biogeochemical-sediment model tailored to Australia's Great
287 Barrier Reef to investigate to what extent realistic OAE applied along a shipping line could
288 alleviate anthropogenic ocean acidification on the reef. Wang et al. (2023) used a coupled ice-
289 circulation-biogeochemical model of the Bering Sea to study the efficiency of OAE in coastal
290 Alaska.

291
292 Implementation of a regional model in a target domain requires generation of a grid with
293 associated bathymetry, specification of boundary conditions (including atmospheric forcing,
294 information about ocean dynamics along the lateral boundaries of the domain, any fluxes of
295 biogeochemical properties across the air-sea, sediment-water, and land-ocean boundaries, river
296 inputs), and generation of initial conditions within the domain (Fennel et al. 2022). Different
297 circulation models are available for implementation in domains targeted for OAE studies (see,
298 e.g., Table 1 in Fennel et al. 2022), all with distinct strengths and established user communities.
299 Particularly relevant in the context of studying coastal applications of OAE is a model's ability
300 to accurately represent coastal topography, making unstructured grid models and models with
301 terrain-following coordinates particularly attractive. Another feature to be considered is a
302 model's ability to run in two-way nested configurations. In the more widely applied one-way
303 nesting of domains, simulated conditions from a larger scale model (referred to as the parent
304 model) are used to generate the dynamic lateral boundary conditions of a smaller scale, higher
305 resolution model (the child model), which runs off line from the parent model. With two-way
306 nesting, both models run simultaneously and information is exchanged continually along their
307 intersecting boundaries. This allows information generated within the high-resolution child
308 domain (e.g., the spreading distribution of a tracer or alkalinity addition) to be received and
309 propagated by the larger-scale parent model. In this context, model simulations are particularly
310 useful if available in near-real time or in forecast mode. This requires specification of lateral
311 boundary conditions and atmospheric forcing up to the present and into the future. Global
312 1/12th-degree nowcasts and 10-day forecasts of ocean conditions are available from the
313 Copernicus Marine Service (CMEMS 2023) and atmospheric forcing up to the present and 10
314 days into the future are available from the European Centre for Medium Range Weather
315 Forecasts (ECMWF 2023).

316
317 One example of a high-resolution local scale model with two-way nested domains is a
318 framework developed for Bedford Basin in Halifax, Canada (Figure 2, Laurent et al. 2024). The
319 model framework consists of three nested ROMS models (ROMS is the Regional Ocean
320 Modelling System; <https://myroms.org>, Haidvogel et al. 2008, Shchepetkin and McWilliams
321 2005). The outermost ROMS domain has a resolution of 900 m and is nested one-way within the
322 data-assimilative global GLORYS reanalysis of physical and biogeochemical properties
323 (Lellouche et al. 2021). Nested within are two models with increasingly higher resolutions of
324 200 m and 60 m. Depending on the scientific objective to be addressed, the models can be run in
325 one-way and two-way nested mode, where two-way nesting is computationally more
326 demanding, and in hindcast or forecast mode. Implementation of dye-tracers within the model
327 (Wang et al. 2024) allows one to determine dynamic distribution patterns and residence times.

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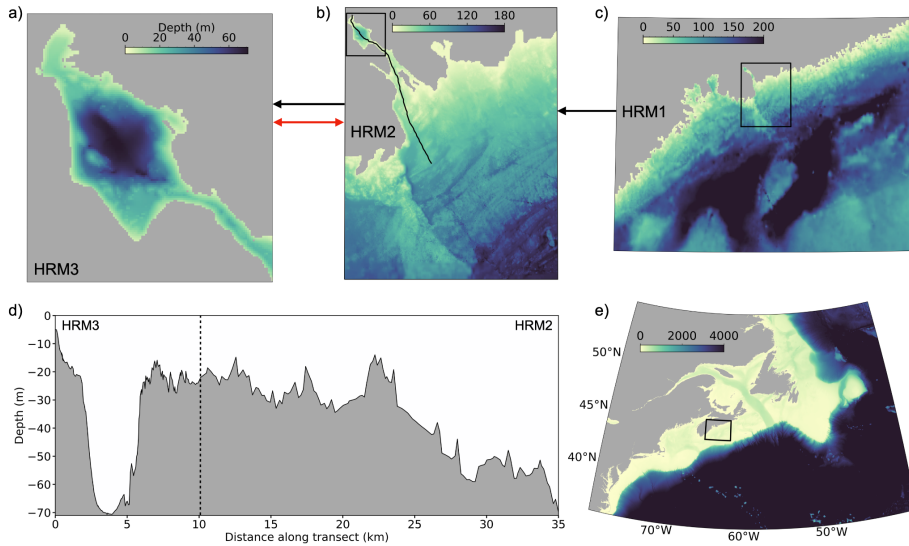


Figure 2: Nested configuration of three ROMS models for the Bedford Basin and the adjacent harbor in Halifax Regional Municipality (HRM). a) The highest resolution model (HRM3; 60 m) includes the 7 km-long and 3 km-wide Bedford Basin and The Narrows, a 20-m shallow narrow channel that connects the basin to the outer harbor. b) The larger scale model (HRM2, 200 m) includes Bedford Basin and Halifax Harbor as well as the adjacent shelf. c) The largest-scale model (HRM3, 900 m) covers the central part of the Scotian Shelf as indicated in e). d) bathymetry along a section through HRM3 and HRM2, indicated by the black line in b). Lateral boundaries of HRM3, HRM2, and HRM1 are shown by black boxes in b), c) and e), respectively. Black arrows indicate the information flow between models in one way nesting mode. The red arrow indicates that HRM1 and HRM2 can be run simultaneously with bi-directional flow of information (two-way coupled mode).

2.1.3. The global scale

A strength of global ocean models is their capacity to comprehensively represent the global overturning circulation and ocean ventilation. These processes control the time scales over which waters are sequestered in the ocean interior and determine how long surface waters are exposed to the atmosphere and can exchange properties, including CO_2 , before being injected back into the ocean interior (Naveira Garabato et al. 2017). Similarly, the large-scale overturning circulation and the patterns associated with ventilation are important to consider in the context of deploying OAE at scale, as these patterns exert strong control on the efficiency of OAE at sequestering CO_2 (e.g., Burt et al. 2021).

354 When global ocean models are dynamically coupled with models of the land biosphere and the
355 atmosphere, they are referred to as Earth System Models (ESMs) and can be employed to
356 explore Earth system feedbacks to mCDR. In the case of OAE, the main feedback is the change
357 in atmospheric $p\text{CO}_2$ and air-sea gas exchange that will result when CDR approaches are
358 implemented at scale. While regional models have to be forced by atmospheric CO_2
359 concentrations, ESMs represent the atmospheric reservoir and are forced by CO_2 emissions into
360 the atmosphere, which then interacts with land and ocean carbon reservoirs. Only the latter
361 approach can account for OAE-induced reductions in the atmospheric CO_2 inventory which, in
362 turn, would lead to a systematic reduction in air-sea CO_2 fluxes. Regional models and global
363 ocean models that do not explicitly represent the atmospheric CO_2 reservoir and instead are
364 forced by prescribed atmospheric $p\text{CO}_2$ cannot simulate the decline in atmospheric $p\text{CO}_2$ due to
365 OAE. Depending on the alkaline material applied, there may also be feedbacks associated with
366 changes in temperature, albedo, nutrient cycles, and biological responses which can be studied
367 with the help of ESMs.

368 Another important strength of global models relates to the fact that anomalies in air-sea CO_2
369 flux generated by OAE deployments will manifest over large spatio-temporal scales because
370 CO_2 equilibrates with the atmosphere via gas exchange slowly. Alkalinity enhanced waters can
371 be transported far away from injection sites before equilibration is complete (He and Tyka
372 2023). Consequently, OAE signals may exit the finite domain of regional models prior to full
373 equilibration with the atmosphere (e.g., Wang et al. 2023). Because global models represent the
374 entire ocean and can be integrated for centuries and longer, they enable full-scale assessments.

375 A primary challenge for global models, however, is that their horizontal resolution is
376 necessarily limited by computational constraints (see example in Figure 3). Most of the global
377 ocean models contributing the Coupled Model Intercomparison Project version 6 (CMIP6), for
378 example, have horizontal resolutions of about 1° or roughly 100 km (Heuzé 2021) and do not
379 accurately represent biogeochemical processes along ocean margins (Laurent et al. 2021). Model
380 grid-spacing imposes a limit on the dynamical scales that can be explicitly resolved in the
381 models; this is particularly problematic for coarse resolution global models because mesoscale
382 eddies—i.e., motions on scales of about 10–100 km—dominate the variability in ocean flows
383 (Stammer 1997). Since coarse resolution models cannot resolve mesoscale eddies explicitly, the
384 rectified effects of these phenomena, including their role in transporting buoyancy and
385 biogeochemical tracers, must be approximated with parameterizations (e.g., Gent and
386 McWilliams 1990).

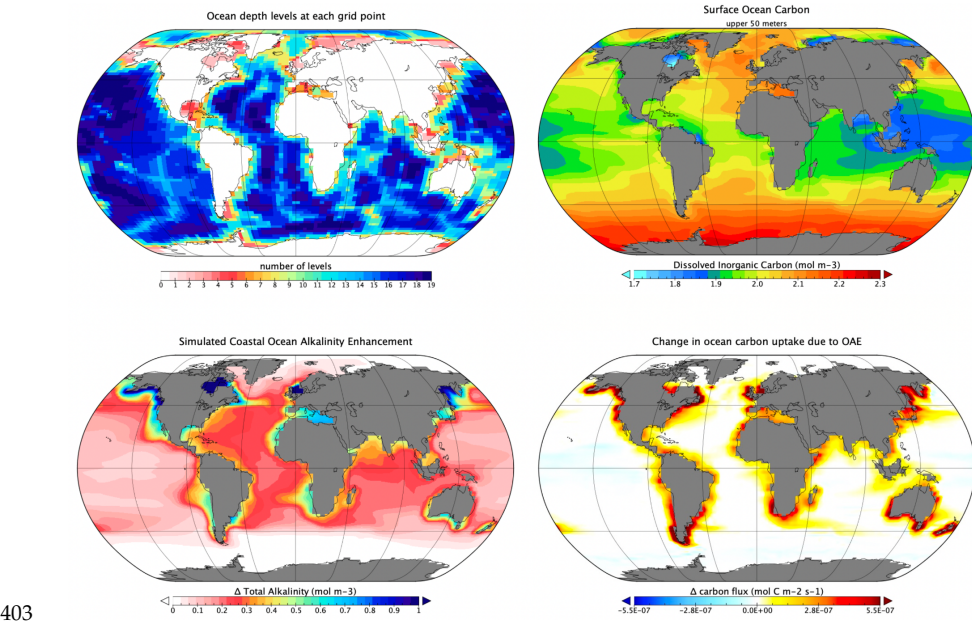
387 Notably, the fidelity of the simulated flow in global models, including the imperfect nature of
388 these parameterizations, projects strongly on the model's capacity to accurately simulate
389 ventilation and the associated uptake of transient tracers, such as anthropogenic CO_2 or
390 chlorofluorocarbons (CFCs), from the atmosphere (e.g., Long et al. 2021). Biases in the uptake of
391 transient tracers will also have implications for a model's capacity to faithfully represent the
392 impact of OAE, where the path of alkalinity-enhanced waters parcels in the surface ocean, and
393 their subsequent transport to depth is a key control on the efficiency of carbon removal. Biases
394 in the simulated flow are also an important determinant of the simulated distribution of

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399 biogeochemical tracers in the model’s mean state. Hinrichs et al. (2023), for example,
 400 demonstrate that inaccuracies in the physical redistribution of alkalinity by the flow is a
 401 dominant mechanism contributing to biases in the alkalinity distributions simulated by CMIP6
 402 models.



403 **Figure 3:** Example of Earth System Model properties and output from the University of Victoria
 404 Earth System Climate Model (Keller et al., 2012, Mengis et al., 2021) including a) the model
 405 bathymetry (depth levels), and b) the simulated present-day dissolved inorganic carbon
 406 concentration (mol m⁻³) averaged over the upper 50 m of the ocean. Panels c) and d) show
 407 results from a coastal OAE study by Feng et al. (2017) where the change in upper ocean
 408 alkalinity (upper 50 m) and the air-sea flux of CO₂ are shown relative to the RCP8.5 control
 409 simulation. Shown is the Oliv100_Omega3.4 simulation from Feng et al. (2017), where 100 μm
 410 olivine grains were added to ice-free coastal grid cells in proportion to RCP 8.5 CO₂ emissions
 411 (i.e., 1 mol of alkalinity per mole of emitted CO₂) until a sea surface aragonite Ω threshold of 3.4
 412 was reached.

414 Finally, another important challenge associated with global ocean models is the requirement to
 415 represent the entire global ocean ecosystem with a single set of model parameters (e.g., Long et
 416 al. 2021, Sauerland et al. 2020). In particular, the biological pump is an important control on the
 417 distribution of biogeochemical tracers, including alkalinity and DIC. The magnitude of organic
 418 carbon export, and the magnitude of biogenic calcium carbonate export, are important controls
 419 on the distribution of alkalinity and DIC at the ocean surface and in the interior (e.g., Fry et al.,

2015). These quantities are a product of ecosystem function and, since the global ocean is characterized by diverse biogeography (e.g., Barton et al., 2013), capturing global variations in the biological pump presents a challenge.

2.1.4 Integration across scales

Choosing the appropriate modeling tool for a given OAE-related question requires clarity about the scale of the problem to be addressed and the objectives of the model application. Approaches for OAE vary significantly with respect to the spatial footprint of alkalinity increase. Proposed methods for spreading alkalinity feedstocks at the surface ocean include the addition of reactive minerals (e.g., CaO, Ca(OH)₂ or Mg(OH)₂) in ship-propeller washes (e.g., Köhler et al., 2013, Renforth et al., 2017, Caserini et al., 2021) or using other means (e.g., Gentile et al., 2022) along tracks from commercial or dedicated OAE vessels or through coastal outfalls (e.g., wastewater-treatment or power plants); the addition of less-reactive minerals to corrosive or high-weathering environments (e.g., olivine spreading on beaches or mineral addition to riverine discharge, e.g., Montserrat et al., 2017, Foteinis et al., 2023, Mu et al., 2023); and electrochemically generated point-sources of alkalinity that are discharged as highly alkaline seawater (e.g., House et al., 2009) from existing facilities (e.g., desalination and wastewater-treatment plants), dedicated facilities (e.g., Wang et al., 2023), or from an array of smaller infrastructure (e.g., grids of off-shore wind turbines). Models for OAE research should represent these footprints of alkalinity increases appropriately for the questions being addressed.

There are research questions that fall relatively neatly into one of the three scale ranges described above in sections 2.1.1 to 2.1.3. For example, consideration of the nearfield effects of different alkalinity feedstocks (e.g., dissolved versus particles) or analysis of the potential impacts from secondary CaCO₃ precipitation due to elevated alkalinity from a point source require models that resolve the scales of turbulent motion. Examination of the change in air-sea CO₂ flux due to a broad and diffuse alkalinity increase is less demanding on model resolution and regional scale models are appropriate for this question. Investigation of Earth system feedbacks requires ESMs. However, there also are many aspects of OAE that require a bridging of scales. For example, when considering different deployment methods like discharge from vessels into the ocean surface boundary layer versus additions made through outfalls via surface or subsurface plumes, modeling requirements vary. In both cases, the resulting biogeochemical response may be affected by dynamics operating in the nearfield, where conditions are sensitive to the deployment method and turbulence has to be considered, and the far-field, where conditions do not depend on the details of how the alkalinity was added and the air-sea flux of CO₂ is instead determined by ambient environmental processes. Another example is the challenge that anomalies in air-sea CO₂ flux generated by OAE deployments will manifest over large spatio-temporal scales because CO₂ equilibrates with the atmosphere via gas exchange slowly. Some interplay among the modeling tools described in sections 2.1.1 and 2.1.2 is likely going to be required. One straightforward approach would be to parameterize small-scale processes in the larger-scale models.

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472 2.2 The range of biogeochemical realism & complexity

473 Application of biogeochemical ocean models for the purposes of OAE research and verification
474 requires reevaluation, and likely further development, of several model assumptions and
475 features related to biogeochemical realism and complexity. For example, the internal sources
476 and sinks of alkalinity are typically not explicitly represented in ocean models; this may become
477 necessary in some circumstances but will be challenging (Section 2.2.1). OAE-related
478 perturbations of alkalinity, other carbonate system properties, and addition of macro- and
479 micronutrients contained in some alkalinity feedstocks may result in biological and ecosystem
480 responses that current biogeochemical models are not capable of representing but that would be
481 relevant for the assessment of environmental impacts of OAE and the verification its CDR
482 efficiency (Section 2.2.2). Furthermore, depending on the environmental setting, sediments can
483 be sources or sinks of alkalinity; these sediment-water fluxes need to be appropriately
484 considered, including the potential impacts of OAE on their magnitude, in order to obtain
485 complete and trustworthy carbon budgets (Section 2.2.3). Other boundary fluxes that require
486 accurate specification are alkalinity inputs from rivers and groundwater (Section 2.2.4) and the
487 air-sea flux of CO₂ across the air-sea interface (Section 2.2.5).

489 2.2.1 Representing alkalinity in seawater

490 Alkalinity is an emergent property that depends on the concentrations of numerous chemical
491 species with distinct internal source and sinks (Schulz et al. 2023; Wolf-Gladrow et al. 2007;
492 Middelburg et al. 2020). Skillful simulation of alkalinity in seawater may require explicit
493 representation of its multiple biotic and abiotic sources and sinks, some of which are difficult to
494 constrain. A major process by which alkalinity is consumed is the production of calcium
495 carbonate. In the water column, this is predominantly a biotic process, performed by calcifiers,
496 although “whiting” events, where calcium carbonate precipitates spontaneously from in
497 ambient seawater can be locally important (e.g., Long et al. 2017).

498
499 Models vary in the degree of mechanistic sophistication with which biogenic calcification is
500 represented. For example, some models explicitly resolve calcifiers, such as pelagic
501 coccolithophores (e.g., Krumhardt et al. 2017) and foraminifera (Grigoratou et al. 2022) and, in
502 some cases, also benthic corals, foraminifera, or calcifying higher trophic levels and thus can
503 mechanistically account for the associated alkalinity consumption. Alternatively, models can
504 parameterize biotic production of carbonate, and its subsequent sinking and dissolution, as a
505 fraction of organic matter production combined with an assumed remineralization profile (e.g.,
506 Schmittner et al. 2008; Long et al. 2021). Dissolution of carbonate minerals produces alkalinity,
507 at the sediment surface and in the water column as carbonate particles sink. This can be
508 represented with first-order abiotic dissolution kinetics with a dependence on the saturation
509 state of ambient water in the water column (e.g., Sulpis et al., 2021), in the sediments (e.g.,
510 Emerson & Archer, 1990) or in micro-environments in aggregates or organisms (Barrett et al.,
511 2014) with systematic differences for different crystal structures, aragonite and calcite (Morse et
512 al., 1980).

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Production of alkalinity occurs via uptake of nitrate or nitrite by photoautotrophs, while remineralization consumes alkalinity when happening aerobically but generates alkalinity when occurring anaerobically, e.g. via denitrification (Fennel et al. 2008). Biotic production and consumption of alkalinity is stoichiometrically coupled to the release or uptake of nutrients and carbon, where non-Redfield processes such as nitrogen fixation or denitrification need to be specifically considered in the stoichiometric relationships (Paulmier et al., 2009).

Spontaneous precipitation of carbonate minerals in pelagic environments could occur when seawater is highly oversaturated with respect to carbonate (Moras et al. 2022) but is, to the best of our knowledge, not yet included in ocean models. When simulating OAE approaches that may generate high oversaturation with respect to carbonate, spontaneous precipitation of carbonates needs to be considered, especially when condensation nuclei are present. Appropriate approaches will have to be developed, e.g., using near-field models to mechanistically represent this process and a meta-model approach to develop parameterizations that are suitable for far-field and larger-scale models.

Organic compounds produced within the ocean or originating from land can also act as proton acceptors and contribute organic alkalinity (e.g., Koeve and Oschlies 2012, Ko et al. 2016, Middelburg et al. 2020) and will impact the carbonate system, the partial pressure of CO₂ and thus the air-sea CO₂ flux. Commonly, the contribution of organic alkalinity is deemed small enough in oceanic environments to be negligible, but this assumption should be reconsidered in the context of OAE, especially for coastal CDR deployments where the organic contribution to alkalinity is thought to be larger. To the best of our knowledge, models do not account for organic alkalinity. A better quantitative understanding of organic contributions to alkalinity is likely needed to parameterize or mechanistically represent its contribution in models. Similarly, it may be important in the context of mineral OAE deployments to account for local variations in [Ca²⁺] and [Mg²⁺] to accurately estimate the pCO₂ anomalies generated by different OAE feedstocks. While these constituents have very long residence times in the ocean, and are hence commonly assumed to vary conservatively in proportion to salinity, variations in their relative abundance has an impact on the thermodynamic equilibrium coefficients used to solve seawater carbonate chemistry (Hain et al., 2015).

2.2.2 Representing biological and ecological processes

A key question related to OAE is whether changes in carbonate chemistry induce differential responses in organisms. In the pelagic zone, OAE might shift the phytoplankton community composition, for example, due to distinct physiological sensitivities of different groups (e.g., Ferderer et al. 2022). Further, if OAE is accomplished via rock dissolution, carbonate versus silicate rock may impact the relative balance between phytoplankton functional groups (PFTs) such as calcifiers and diatoms, and changes in Mg and Ca ratios may also influence calcification (Bach et al., 2019). Additionally, ancillary constituents specific to particular feedstocks may have biological activity. Silicate rocks include bioreactive metals such as Fe, a micronutrient with the capacity to stimulate phytoplankton growth, and others that are can be toxic when occurring in

high concentrations, such as Ni and Cu, and may adversely impact phytoplankton and reduce primary productivity (Bach et al., 2019). The bioreactivity of these metals may be difficult to simulate in models as their dissolved concentrations can be partially mediated by complexation with organic ligands (Guo et al., 2022). Physical impacts of OAE feedstocks may also have important biological impacts through changes in the propagation of light in the surface ocean, and direct exposure to mineral particles may have additional impacts, e.g., on zooplankton through particle ingestion (Harvey, 2008; Fakhraee et al., 2023). Effects of OAE on plankton have the potential to propagate to higher trophic levels through marine food webs as the magnitude and quality of net primary productivity shifts and trophic energy transfer is altered accordingly.

Simulating this full collection of processes in models is challenging. Dominant modeling paradigms for simulating planktonic ecosystems include PFT- and trait-based models (e.g., Negrete-Garcia et al., 2022). In these systems, physiological sensitivities are parameterized according to transfer functions that modulate rate processes—growth, for instance—on the basis of ambient environmental conditions. Nutrient limitation of growth is often represented using Michaelis–Menten kinetics wherein growth rates decline as nutrients concentrations become limiting. State-of-the-art ESMs represent PFTs with multiple nutrient co-limitation, which is essential to effectively simulate plankton biogeography of the global ocean. Diatoms, for example, are capable of high growth rates, enabling them to outcompete other phytoplankton under high-nutrient conditions, but their range is restricted to high latitudes and upwelling regions where there is sufficient silicate. If OAE were to modulate the concentration of constituents represented by multiple nutrient co-limitation models, it is possible such models could simulate the phytoplankton community response—though it’s important to consider whether the models provide representations that are sufficiently robust for the magnitude of OAE-related perturbations. In some cases, models are missing key processes that would be required to mechanistically simulate certain effects. We are aware of no models that represent Ni toxicity, for instance. Including these effects, as well as a capacity to simulate secondary interactions, such as ligand complexation of metals in OAE feedstocks, will require significant investment in empirical experimentation to understand essential rate processes and physiological responses.

Shortcomings in the capacity of models to represent physiological responses to OAE is an important consideration for the ability of models to faithfully represent ecological impacts. Notably, electrochemical OAE techniques present a simpler set of processes to consider than using crushed-rock feedstocks, where ancillary constituents and physical dynamics come into play. For electrochemical OAE, the most likely biological feedback to consider relates to the impacts of changing carbonate chemistry on biogenic rates of calcification or phytoplankton growth rates (Paul and Bach 2020). It is also possible that carbon limitation of phytoplankton growth (Paul and Bach 2020; Riebesell et al. 1993) may also be important. Empirical research exploring physiological sensitivities should be used to develop prioritizations of key model processes comprising early targets for implementation. Model documentations should use consistent stoichiometric relations to link alkalinity changes to those of nutrients and carbon

605 (Paulmier et al. 2009) and state the assumptions made about carbonate formation and
606 dissolution.

608 2.2.3 Representing sediment-water exchanges

609 The exchange of solutes between the sediments and overlying water influences ocean
610 chemistry, including the properties of the carbonate system (Burdige 2007). Depending on
611 location and time scale, OAE may affect these exchanges and should be appropriately
612 considered in models. Sediments influence the marine carbonate system primarily through the
613 remineralization of organic matter, which returns DIC to overlying water (and alkalinity if this
614 remineralization occurs anaerobically), and the dissolution of biogenic silicate or carbonate
615 minerals. CaCO_3 is of particular importance as its dissolution releases alkalinity, while its burial
616 is an alkalinity sink, and the balance between the two is a key control on the ocean's alkalinity
617 balance over timescales approaching 10^4 years (Middelburg et al. 2020). Furthermore,
618 remineralization and other microbial metabolisms, such as "cable bacteria," can significantly
619 lower pore water pH by several pH units below seawater values (Meysman and Montserrat
620 2017). This can drive dissolution of CaCO_3 and generate alkalinity in the sediments, even in
621 shallow waters when the overlying water is supersaturated (Rau et al. 2012).

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622 Representing these processes in coastal and shelf sediments (< 200 m) is challenging. Shallow
623 water depths and high productivity result in a significant delivery of organic matter to the
624 sediments that is much larger than in the deep ocean. As a result, the relative importance of
625 sediments in organic matter remineralization is larger and production of alkalinity by anaerobic
626 metabolisms is more important in these shallow sediments than in the deep ocean (Seitzinger et
627 al. 2006, Jahnke 2010, Huettel et al. 2014, Chua et al. 2022). In addition, these environments are
628 dynamic with organic supply and bottom water conditions varying on tidal, seasonal, and
629 interannual timescales. Accounting for the exchange between sediments and overlying water
630 and its variability on tidal, seasonal, and interannual timescales will likely be necessary in
631 regional and global biogeochemical models that aim to simulate alkalinity cycling in coastal and
632 shelf seas, even for relatively short simulation durations of months to years.

634 The choice of approach to modeling sediments may depend on the sediment type. For example,
635 the mechanisms transporting solutes across the sediment-water interface can be divided into
636 two categories depending on the sediment's grain size. In coarse sediments, i.e. permeable
637 sands, pressure gradients drive flow through the seabed replenishing sediment oxygen content
638 (Huettel et al. 2014). Organic carbon stores are low and remineralization was long thought to be
639 primarily aerobic. However, evidence has emerged relatively recently that anaerobic
640 remineralization in sandy sediments is more important than originally thought (Chua et al. 2022
641 and references therein). Idealized models that represent the three-dimensional sediment
642 structure illustrate the importance of turbulence and oscillatory flows in permeable sediments
643 (see Box 2 in Chua et al. 2022). These models are highly localized and computationally
644 demanding, prohibiting their coupling with ocean biogeochemical models. Thus, permeable
645

sediments are currently not well represented in regional or global ocean biogeochemical models.

In cohesive, fine-grained sediments with low permeability, i.e. muds, transport is limited by diffusion or faunal mediated mixing and exchange processes, i.e. bioirrigation or bioturbation (Meysman, et al. 2006, Aller 2001). In these environments, detailed multicomponent reactive-transport models of sediment biogeochemistry – so called diagenetic models – can reproduce carbon remineralization rates partitioned between aerobic and anaerobic pathways, precipitation/dissolution reactions between sediment grains and porewaters, and the transport of solutes across the sediment-water interface (Boudreau 1997, Middelburg et al., 2020). These mechanistic models will be useful for detailed investigations into how perturbations of the carbonate system in seawater overlying the sediments affect their biogeochemistry and for addressing questions about the potential influence of particulate alkalinity feedstocks settling to the seafloor (Montserrat et al. 2017, Meysman and Montserrat 2017). However, typically these models are one-dimensional and applied to a few representative locations. Coupling fully explicit diagenetic models to three-dimensional ocean biogeochemical models, while conceptually straightforward, is computationally prohibitive. Instead, depth-integrated sediment processes have been implemented as bottom boundary conditions (e.g., Moriarty et al. 2017, 2018, Laurent et al. 2016). For example, Laurent et al. (2016) used a diagenetic model in a “meta-modeling” approach to estimate bottom boundary nutrient fluxes for a regional scale biogeochemical model. By parameterizing the diagenetic model with detailed geochemical data (porewater profiles and nutrient fluxes) from a few individual locations, then forcing it over a range of expected bottom water conditions, they developed empirical functions relating sediment fluxes to bottom water conditions that could be used to parameterize bottom boundary conditions in the water column model. A similar approach could be used in OAE models to parameterize how sediment biogeochemistry may alter alkalinity fluxes, for example, how redox sensitive processes, such as coupled nitrification-denitrification or sulfate reduction coupled to pyrite burial, both of which may produce alkalinity (Soetaert et al. 2007), may respond to changes in bottom water oxygen or organic matter loading.

When considering the long-term storage of CO₂ in global-scale ESMs, the interactions between sediments and the deep ocean (> 1000 m bottom depth) may need to be considered. In this environment most organic matter remineralization occurs in the water column, and the small amount of organic matter reaching the seafloor is remineralized aerobically with little to no release of alkalinity. In this case, sediment remineralization can likely be either ignored or implemented as a reflective boundary condition where the simulated POC flux to the seafloor is immediately returned as DIC and remineralized nutrients. However, the dissolution or preservation of CaCO₃ in deep sediments is critical to controlling deep water alkalinity and may be important in model simulations that aim to quantify OAE effects on the timescales associated with the large-scale global overturning circulation. CaCO₃ solubility increases with pressure and decreasing pH and CaCO₃ eventually becomes undersaturated at depth. The depth at which sinking CaCO₃ balances its dissolution is referred to as the carbonate compensation depth (CCD). An increase in bottom water CO₃²⁻ or CaCO₃ deposition, will deepen the CCD,

burying CaCO_3 , trapping alkalinity, and lowering the alkalinity budget of the ocean. Conversely if CaCO_3 rain rate or CO_3^{2-} concentration decreases, the CCD will shoal and previously buried CaCO_3 will dissolve releasing alkalinity to the deep ocean. CCD compensation therefore opposes any forcing of the deep ocean carbonate system and therefore dampens the rise of CO_2 in the atmosphere but will also counteract any potential OAE solution (see Renforth and Henderson 2017 for a detailed explanation). Although most CaCO_3 dissolution occurs in the sediments, there is no consensus as to the level of detail this needs to be represented in models. Some global models employed to investigate large-scale OAE include calcium carbonate dynamics at the sediment surface (Ilyina et al. 2013) others disregard this process (Keller et al. 2014).

Often global models will parameterize CaCO_3 burial as a function of saturation state, such an approach is effective for resolving CCD dynamics over geological timescales ($\sim 10,000$ y), but not over the century to millennial timescales of CCD readjustment. Models that fully couple sediment diagenesis can resolve these dynamics (Gehlen et al. 2008), but the computational demand can make them ineffective. One solution is the approach of Boudreau et al. (2010) and (2018). By suggesting that CaCO_3 dissolution dynamics are controlled by transport of dissolution products across the benthic boundary layer, they were able to derive equations predicting CCD depth and CaCO_3 dissolution based on bottom water CO_3^{2-} and CaCO_3 rain rate and avoiding a detailed representation of the sediments. These equations, combined with model bathymetry, can parameterize sediment CO_3^{2-} flux as a boundary condition and suitably account for transient sediment CaCO_3 dissolution in large scale ESMs while avoiding the computational demands of a fully coupled ocean circulation-diagenesis model.

2.2.4 Representing river and groundwater fluxes

Regional and global ocean biogeochemical models typically account for river inputs, including their contributions to alkalinity and **PIC**. In most models this is done by specifying alkalinity and **PIC** concentrations in imposed riverine freshwater fluxes, although accurate prescription of these concentrations can be challenging. Typically, a combination of direct river measurements, where available, output from watershed models (e.g., Seitzinger et al. 2010), or extrapolations of coastal ocean measurements to a freshwater endmember (e.g., Rutherford et al. 2021) are used. Solute inputs from groundwater are typically ignored but could be important locally. In high-resolution coastal domains near urban areas, sewage input may be an additional important source of carbon, nutrients, and alkalinity.

It is important to note that land-based CDR applications may have an important effect on ocean alkalinity dynamics through riverine and groundwater delivery of solutes. Terrestrial OAE equivalents broadly referred to as Enhanced Rock Weathering (ERW) rely on the application of lime or pulverized silicate or carbonate rocks on land and in rivers. These strategies aim to generate CO_2 uptake locally but yield a leaching flux of bicarbonate into freshwater systems and subsequent transport into the coastal ocean. Field trials and some commercial applications are currently underway, most of them with the implicit or explicit assumption that the enhanced

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delivery of alkalinity will generate a carbon removal in the ocean (Köhler et al., 2010; Taylor et al., 2016; Bach et al., 2019). There is a need for coordinated efforts to improve quantification of background riverine fluxes and establish initiatives to effectively track the solute additions from ERW.

2.2.5 Representing air-sea gas exchange

The calculation of air-sea gas exchange is necessary for the quantification of net carbon uptake from OAE in models. Biogeochemical models typically represent this exchange using a bulk relationship that depends on the product of the gas transfer velocity and the effective air-sea concentration difference (Fairall et al. 2000). However, the gas transfer velocity remains highly uncertain and is sensitive to a collection of processes that vary across scales, including sea state, boundary layer turbulence, bubble dynamics, and concentrations of surfactants. The most widely used parameterizations of the gas transfer velocity use empirical fits to observations to construct a functional relation dependent on wind speed only, under the premise that turbulence and bubbles (via the breaking of surface gravity waves) are predominantly determined by wind stress (Wanninkhof 2014). This neglects processes that could be regionally important such as convection, modification by biological surfactants, rain and wave-current interactions, while vastly simplifying the effects of wave breaking and bubbles. Although different dependencies on wind speed have been proposed (quadratic, cubic, hybrid), parameterizing the gas transfer coefficient as a quadratic function of the 10-meter wind speed is the most common (Wanninkhof 1992; Wanninkhof 2014). This relationship is supported by direct measurements of air-sea flux at intermediate wind speeds (3-15 m/s), but at low wind speeds (< 3 m/s), non-wind effects can have an important impact on gas transfer. At high wind speeds (> 15 m/s), breaking waves and bubble injection enhance gas exchange for lower solubility gasses such as CO₂ (Bell et al. 2017). Therefore, quadratic fits tend to underestimate the gas exchange at low and high wind speeds (Bell et al. 2017).

More complex air-sea exchange parameterizations account for processes such as bubbles, near surface gradients and buoyancy driven convection (e.g., Liang et al. 2013, Fairall et al. 2000), but they depend upon a wider range of input variables. Other considerations in estimating flux arise from the nonlinear dependence on these variables, e.g., wind speed, which can lead to underestimates when made using daily averages rather than hourly measurements (Bates and Merlivat 2001).

Notably, the gas transfer velocity (k_w) determines the kinetics of gas exchange, given a perturbation in surface ocean $p\text{CO}_2$ away from equilibrium. The timescale for CO₂ equilibration over the surface mixed layer can be fully quantified using the following expression,

$$\tau_{gas-ex} = \left(\frac{\partial \text{CO}_2}{\partial \text{DIC}} \right)^{-1} \left(\frac{h}{k_w} \right)$$

where h is the depth of the surface mixed layer and the partial derivative $\partial \text{CO}_2 / \partial \text{DIC}$ captures the thermodynamic state of the carbon system chemistry in seawater, specifically with respect to the amount that dissolved CO₂ changes per unit change in DIC (Sarmiento and Gruber 2006). This property is related to the buffer capacity and varies in roughly linear proportion to the

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779 carbonate ion concentration. The magnitude of $\left(\frac{\partial \text{CO}_2}{\partial \text{DIC}}\right)^{-1}$ is typically about 20, which explains
780 why the equilibration timescale for CO₂ is so long. The contribution of uncertainty in the gas
781 exchange velocity to overall uncertainty in carbon uptake from OAE deployments will depend
782 in part on the circulation regime involved. For example, in situations where alkalinity-enhanced
783 water parcels are retained at the surface for timescales that are significantly longer than $\tau_{\text{gas-ex}}$,
784 full equilibration will occur and the impact of uncertainty in the gas exchange velocity will have
785 limited influence on the overall uncertainty.

786
787 Even though OAE-induced additional air-sea CO₂ fluxes will, even in hypothetical massive
788 deployments, amount to at most a few Gt CO₂/yr, which is typically not more than a percent of
789 the atmospheric CO₂ inventory, this subtle difference in the treatment of the atmospheric
790 boundary condition can be significant. Using prescribed atmospheric $p\text{CO}_2$ that is unresponsive
791 to marine CDR-induced air-sea CO₂ fluxes has been shown to overestimate oceanic CO₂ uptake
792 by 2%, 25%, 100% and more than 500% on annual, decadal, centennial, and millennial
793 timescales, respectively (Oschlies, 2009). Simulations with prescribed atmospheric $p\text{CO}_2$ need to
794 take such systematic biases into account.

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796 2.3 Model development needs for OAE research

797 While there is already substantial capacity for simulating ocean biogeochemical dynamics at
798 global to regional scales, the discussion above implicates several areas where additional efforts
799 are required to fully establish a modeling capability suitable for supporting OAE. These fall into
800 four primary areas: (1) supporting multi-scale simulations with sufficiently high-fidelity flow
801 fields; (2) faithfully simulating the near-field dynamics associated with alkalinity addition; (3)
802 capturing feedbacks to OAE owing to biological and geochemical responses; and (4) identifying
803 whether there are reduced-complexity modeling approaches that might provide sufficiently
804 robust estimates of the net effects of OAE.

805 As elucidated above, a primary consideration related to capturing OAE impacts is the fidelity of
806 the simulated flow. Notably, OAE presents a somewhat novel use case requiring an effective
807 multi-scale modeling capability. A conceptually straightforward path to improving the
808 representation of ocean circulation and mixing is to increase the resolution of the model grid.
809 However, the computational demand of high-resolution simulations can only be met over more
810 limited-area domains. Since the spatiotemporal footprint of OAE-related perturbations is likely
811 to be large, there will be a need to represent large regions. An argument might be made,
812 however, that the circulation in proximity of an OAE site is most important to capture with
813 high-fidelity. This can be achieved with two-way nested regional models as described in see
814 Section 2.1.2 but will require further development to couple in the nearfield models described in
815 Section 2.1.1. Native grid-refinement, e.g. via unstructured grids, is another approach that may
816 be pursued to effectively support OAE research.

817 The second area of model development relates to the requirement of faithfully representing the
818 dynamics associated with alkalinity addition. Regional to global scales are the most relevant for
819 simulating the air-to-sea exchange of CO₂ ensuing from OAE. It is important, however, to

822 ensure that local processes affecting the mass fluxes and initial dispersal of alkalinity are
823 handled appropriately. As illustrated above, DNS or LES simulations (section 2.1.1) can be
824 leveraged to develop parameterizations for larger-scale models, including for crushed-rock
825 feedstocks where particle dynamics may be important or techniques involving alkalinity
826 enhanced streams entering the ocean from outfall pipes. In addition to process fidelity, there are
827 also numerical constraints to consider. For example, advection schemes used in most ocean
828 general circulation models struggle to represent sharp gradients; large mass fluxes of alkalinity
829 into single model grid-points are likely to cause advection errors that may contaminate aspects
830 of the model solutions making interpretation difficult. More specifically, conservative advection
831 schemes can be characterized in terms of their accuracy, monotonicity (i.e., ability to preserve
832 sign), and linearity (i.e., ability to preserve additivity) and there are always tradeoffs to make
833 between these properties. Research may be required to determine which schemes are best
834 suited to the particular challenges associated with representing the advection of OAE signals.

835 The third area of model development relates to our capacity to fully capture the range of
836 biogeochemical feedback associated with OAE. The class of processes to consider here is
837 potentially large and many have been touched on in section 2.2.1 to 2.2.3. Precipitation
838 dynamics, specific elemental components of alkalinity, biogenic responses mediated by
839 physiological or ecological sensitivities, impacts and processes controlling the cycling of
840 ancillary constituents, and accurate sediment-water exchange are all areas that merit
841 consideration. Further efforts are required to understand and prioritize these areas of potential
842 development and, notably, their relative importance is likely to be regionally dependent.

843 Finally, it is important that models be tailored to address specific questions of relevance. In this
844 context, it may be important to consider how much model complexity is required to capture the
845 effects of perturbations, seeking parsimonious representations that are well-supported by
846 empirical constraints and invoking wherever possible a separation of concerns to isolate the
847 factors contributing to uncertainty. For example, there are several near-field considerations that
848 might be addressed using a combination of local observations and ultra-high-resolution
849 modeling tools to generate estimates of alkalinity mass fluxes that are subsequently imposed as
850 forcing in regional- to global-scale models. Another key question is how important it is to
851 comprehensively simulate the mean state to faithfully capture the response to OAE
852 perturbations for the purpose of MRV. For example, if it can be documented that biological
853 feedbacks to OAE are of negligible concern, the core target for simulating OAE effects for MRV
854 may be to capture the cumulative integral of air-sea CO₂ exchange associated with the induced
855 surface ocean pCO₂ anomaly. The mean state of the seawater carbon system is relevant here as
856 the background DIC and alkalinity fields determine the pCO₂ response per unit addition of
857 alkalinity, but fully prognostic calculations of nutrient cycling may not be necessary.

858 **3 Model validation and integration with observations**

859 Whether a model is useful for OAE research depends on how accurately it represents the
860 physical, chemical, and biological processes that are relevant to the specific research question to
861 be addressed. Model validation, the evaluation of a model's performance, and estimation of

uncertainties in model output should thus be integral parts of model implementation and application. It is important to note that any model, even after best efforts have been made to improve formulations and conduct the most thorough validation, will deviate from reality. Any model is, by definition, a simplification of the real world and thus its output will be subject to uncertainties. Deviations of the model state from the real world can be reduced by applying statistical techniques, collectively referred to as Data Assimilation (DA) methods, that combine models with observations and yield the best possible estimates. The steps typically involved in model implementation and validation, and possible integration with observations through data assimilation are shown in Figure 3. In this section, we summarize the most important observation needs for model validation (Section 3.1), briefly describe typical metrics for model validation and articulate a reasonable minimum criterion (Section 3.2), give a high-level explanation of approaches for the formal statistical combination of models with observations through parameter optimization and state estimation (Section 3.3), and describe approaches for the specification of uncertainty in model outputs (Section 3.4).

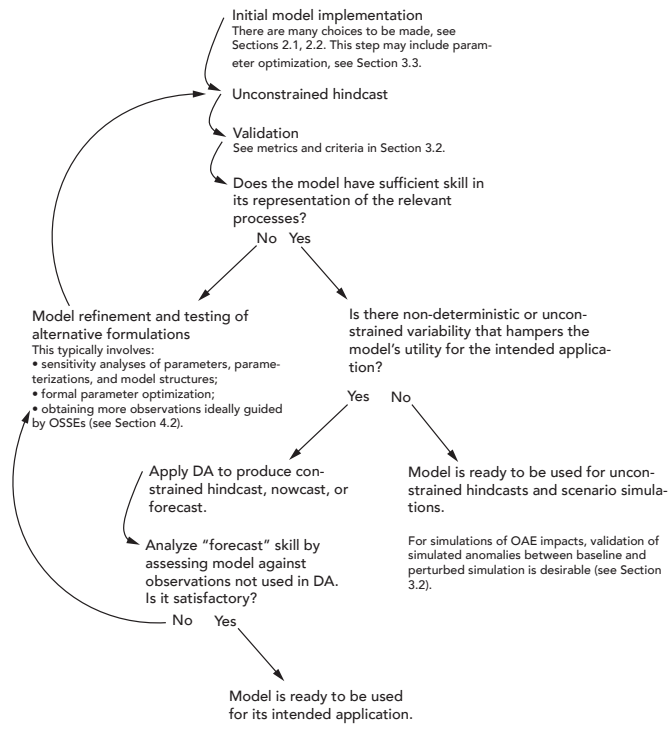


Figure 3: Typical steps in model implementation and validation.

880 3.1 Observation types for validation

881 Two fundamental requirements for models to be useful in the context of OAE research are high-
882 fidelity representations of physical transport due to advection and mixing, and of
883 biogeochemical effects of OAE, most importantly changes in the inorganic carbon properties.

884 Observations for validation of the simulated physical transport of alkalized waters include
885 temperature and salinity distributions, direct measurements of currents, surface drifter
886 trajectories, sea surface height observations from satellite altimetry, and estimates of
887 geostrophic flow derived from the latter. Additional metrics relevant for assessing the fidelity of
888 the large-scale overturning circulation in global models include combinations of biogeochemical
889 concentration and transient tracers. For example, oxygen can be useful for identifying large-
890 scale transport pathways, even though it convolutes dynamical and biological information.
891 Particularly valuable for assessing large-scale ocean transport on the timescales relevant for
892 OAE are abiotic transient tracers such as such as chlorofluorocarbons (CFCs), sulfur
893 hexafluoride (SF₆), and possibly the isotopes ³⁹Ar and ¹⁴C. Observational approaches for
894 validation at regional scales include explicit tracer studies for documenting dispersion
895 properties using Rhodamine dye or SF₆.

896 In addition to the dynamics of the flow, model validation for OAE research requires the
897 assessment of the fidelity of simulated carbonate chemistry variables (e.g., alkalinity, total
898 dissolved inorganic carbon or DIC, pH, $p\text{CO}_2$) and salinity and temperature, which are used to
899 calculate the 13 thermodynamic equilibrium constants and conservative chemical species
900 needed to constrain seawater acid-base chemistry in oxygenated seawater. Depending on the
901 OAE approach and the model application, assessment may also require observed macronutrient
902 (e.g., nitrate, silicate, or phosphate), micronutrient (e.g., Fe), and contaminant (e.g., Ni, and Cr)
903 measurements; bulk seawater properties related to biogeochemical cycling (e.g., dissolved
904 organic carbon content [DOC], particulate inorganic carbon [PIC], chlorophyll fluorescence);
905 and biogeochemical rates and fluxes (e.g., net community calcification).

906 It is not always feasible to obtain the ideal carbonate system observations for model validation.
907 Temperature and salinity can be measured reliably across all ocean depths and, with greater
908 uncertainty and only at the ocean surface, remotely from satellites. The technical capacity for
909 seawater pH measurements is evolving rapidly and sensors and systems now exist for pH
910 measurements across nearly all depths, though the depth-capable systems require regular
911 recalibration (e.g., Maurer et al., 2021). Similarly, there are numerous ways to observe surface
912 ocean $p\text{CO}_2$ using a variety of crewed, autonomous, and fixed-location platforms (e.g., ship-
913 based, Saildrone, and moored systems). However, interior-ocean $p\text{CO}_2$ observations remain
914 challenging to obtain due to the need for calibration gasses and a gas-water interface. Alkalinity
915 titrations are predominantly performed on discrete bottle samples collected by hand, though
916 autonomous titration systems are under development that enable *in situ* surface time series
917 measurements (Shangguan et al., 2022). Microfluidic *in situ* alkalinity titrators are also under
918 development that consume less reagent per sample but currently show higher uncertainties
919 than discrete samples (Sonnichsen et al. 2023). Solid state titrators that generate acid titrant *in*

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situ show promise for surface and subsurface alkalinity titrations, but these sensors are still undergoing development and validation (Briggs et al., 2017). DIC observations combine the limitations of current measurement systems for both the $p\text{CO}_2$ and alkalinity, and there are only a handful of automated DIC titration systems rated for surface ocean measurements (e.g., Fassbender et al. 2015; Wang et al. 2015; Ringham 2022). Theoretically, measurement of two of the carbonate system parameters in combination with temperature and salinity and some additional assumptions allows calculation of the other carbonate system parameters in seawater. Unfortunately, the pair of $p\text{CO}_2$ and pH, which are the most accessible to autonomous measurement among the carbonate system parameters, provide nearly identical information about the system. Thus, the results of the calculations that use this pair have higher uncertainties than other combinations (Dickson and Riley 1979; Millero 2007; Cullison Gray et al. 2011; McLaughlin et al. 2015; Raimondi et al. 2019) and are therefore not ideal as a pair for model validation.

3.2 Validation metrics and approach

Validation relies on comparing the model output to observations, often in an iterative loop where the evaluation of a hindcast simulation is followed by model refinements followed in turn by a new hindcast and re-evaluation (Figure 3, Rothstein et al. 2006). Several evaluation metrics are commonly used (see Box 3 in Fennel et al. 2022). The three most common are the root-mean-square error (RMSE), the bias, and the correlation coefficient. All three are relative measures without any objective criterion that indicates which range of values is acceptable or unacceptable. In contrast, the Z-scores, which consider variability within the observational data set, and the so-called model efficiency or model skill, which quantifies whether the model outperforms an observational climatology are two metrics with built-in criteria as to whether a model's performance is acceptable or not (Fennel et al. 2022). Since no single metric provides a complete picture of a model's skill, multiple complementary metrics should always be used in combination (Stow et al. 2009). Furthermore, different points in space and time, and a breadth of variable types should be part of any comprehensive validation because a model may provide accurate estimates for some variables, locations, or times but perform poorly for others (Doney et al. 2009).

For OAE research, validation can be considered as a two-step challenge. First, it is necessary to validate unperturbed model baselines to gain confidence that the natural variability is represented appropriately and to quantify model uncertainties. One should compare model-simulated spatial fields and time-series at strategic locations with appropriate observations to assess the model's skill at representing mean distributions as well as the variability for carbonate chemistry measurements and other relevant properties using several of the complementary quantitative metrics listed above. A model could be considered as sufficiently validated when mean distributions, their seasonal variability, and the timing and magnitude of events (e.g., blooms, physical disturbances) are accurately represented. As described in Section 3.1, insufficient availability of observational constraints on carbonate system parameters presents a major challenge in this regard. In models applied for OAE research, it is particularly important to assess whether they realistically capture the distributions and variability of

seawater properties that govern sensitivity of the seawater carbonate system; recent work by Hinrichs et al. (2023) shows that the current representation of alkalinity in state-of-the-art models requires improvements.

The second, even more difficult step is to test whether a model accurately represents alkalinity additions. OAE-related modeling studies thus far have relied on models that are validated only for baseline conditions. These are useful as sensitivity studies. However, validation of a model's ability to accurately represent the perturbations of an alkalinity addition is ultimately needed to address OAE science questions around environmental impacts and MRV. It is likely that the metrics described above for baseline validation are not suitable for this task. Validation should focus on quantifying whether the model accurately captures the anomalies created by OAE. This requires consideration of the spatial footprint and temporal evolution of perturbations and ideally a close integration of experimental, observational, and modeling efforts. For example, a model that is deemed skillful after baseline validation can be used to estimate the appropriate dosage of alkalinity additions, thus ensuring a measurable signal, and guide the observational strategy; subsequent validation may indicate model shortcomings that were not obvious in the baseline validation (e.g., diverging dissipation rates between model and field observations) and prompt model refinement in an iterative loop of model validation, improvement, and renewed experimental assessment (Figure 3).

It is important to note that even with repeated steps of validation and model improvement, there is going to be a limit to the degree of realism that can be achieved with any model. Any model simulation will be prone to errors and uncertainties. Sources of error include inaccuracies in model inputs, numerical approximation schemes, insufficient process understanding, and inaccurate model parameters and parameterizations.

3.3 Data Assimilation

Data assimilation (DA) is the process of improving the dynamical behavior of models by statistically combining them with observations. There are a variety of DA techniques that rely on different mathematical and statistical approaches (Carrassi et al. 2018). Originally developed for numerical weather prediction, DA has been successfully applied to ocean models, including biogeochemical models (Mattern et al. 2017, Cossarini et al. 2019, Ciavatta et al. 2018, Verdy and Mazloff 2017, Teruzzi et al. 2018, Fennel et al. 2019) but success critically depends on the information content of the available observations (Yu et al. 2018; Wang et al. 2020). While DA has been shown to yield large improvements in important parameters governing biogeochemical processes (Mattern et al. 2012, Schartau et al. 2017, Wang et al. 2020) and in model estimates of the physical and biogeochemical model state (Hu et al. 2012, Mattern et al. 2017, Ciavatta et al. 2018), it is only starting to be applied to carbonate system properties (Verdy and Mazloff 2017, Carroll et al. 2020, Turner et al. 2023, Figure 4).

Application of DA for ocean models is typically applied for one of two purposes: (1) to systematically optimize model parameters, e.g., phytoplankton growth and nutrient uptake or rates of background dispersion, and (2) to estimate the ocean state, e.g., distributions of temperature, phytoplankton biomass, alkalinity (see Fennel et al. 2022 for more details on the

practical approaches and examples). The first purpose addresses systematic errors and biases in models and is useful when systematically modifying and testing different model formulations while the second assumes an unbiased model and addresses unresolved stochasticity, e.g., correcting the locations of mesoscale eddies and current meanders. State estimation offers the potential to constrain variability such that OAE-induced perturbations of carbonate system parameters can be documented even if they are smaller than the natural variability in the study region. Joint estimation of physical and biogeochemical properties is common and can yield significant improvements for both types of properties (Yu et al. 2018). Hybrid approaches combining parameter and state estimation have also been proposed (Kitagawa 1998, Mattern et al. 2012, 2014) but are less widely used.

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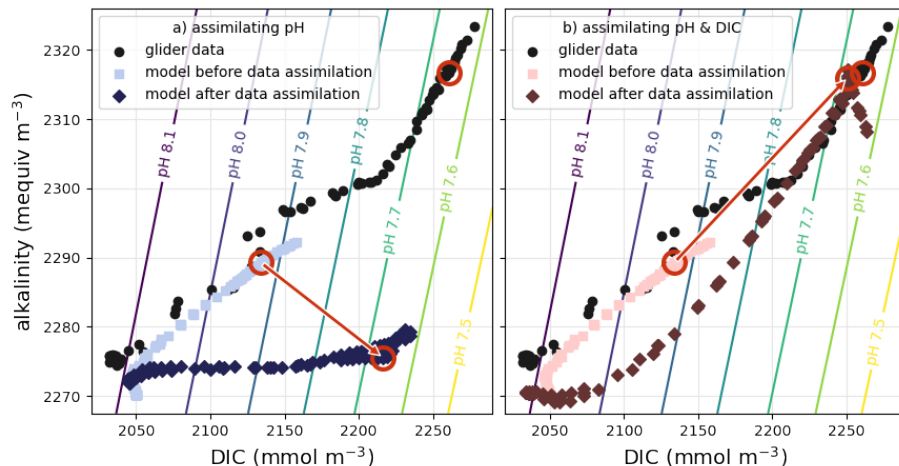


Figure 4: Example of a DA application for state estimation of carbonate system properties within a 3-dimensional model of the California Current System. The symbols show glider data and model estimates at the measurement times and locations; one specific data point and its associated model estimates are highlighted by red circles. Each data point consists of measured pH alongside estimated alkalinity and DIC values (see Takeshita et al. (2021) for data source and details). In the model, pH is a diagnostic variable and primarily dependent on the model's alkalinity and DIC estimates. (a) When only pH data is assimilated, the model estimates are moved closer to the observed pH values by increments in alkalinity-DIC space that degrade the model's alkalinity estimates. (b) The model state estimates improve considerably by assimilating data for DIC (or alkalinity; not shown) together with the pH observations.

Successful application of DA critically requires sufficient observations either of the properties that the model parameters to be estimated depend on or of the state variables that are being estimated. The most commonly used observation type in biogeochemical DA applications is satellite-based ocean color observations (Mattern et al. 2017, Ciavatta et al. 2018, Teruzzi et al.

1031 2018) which are available at a relatively high temporal resolution and covering large areas of the
 1032 surface ocean. While these observations are useful for informing model estimates of properties
 1033 directly linked to processes involving phytoplankton, they provide little information on the
 1034 carbonate system. Dynamical models are able to quantitatively constrain processes that cannot
 1035 be measured directly, by inferring them from observable properties, but only if the observations
 1036 contain enough relevant information about the processes of interest. Hence, one of the biggest
 1037 challenges facing the application of DA to models of the marine carbonate system, is the
 1038 sparsity of observations of the marine carbonate system. Observations of pH, pCO₂, alkalinity,
 1039 and DIC used to be limited to moorings and research cruises but have more recently been
 1040 extended by automated observing systems, such as gliders, BGC-Argo floats and uncrewed
 1041 surface vehicles (Bushinski et al. 2019). Although these measurements are becoming more
 1042 common (Chai et al. 2020), they are still sparse compared to what is typically required for DA
 1043 applications. In this context, an additional challenge is the problem of underdetermination, i.e.
 1044 if multiple processes or properties of interest can cause a similar change in an observable
 1045 property, then observing this property alone may not hold enough information to constrain
 1046 these processes or properties and more observations are needed (see Fig. 4 and code examples
 1047 in Fennel et al. 2022). As new platforms are added to the observing system, DA techniques can
 1048 help guide their optimal deployment and tailor observational programs to the specific needs of
 1049 OAE applications (see Section 4.3 below). Furthermore, statistical and machine-learning
 1050 approaches are being developed (e.g., Lohrenz et al. 2018, Bittig et al. 2018, in prep.) that may
 1051 help overcome the undersampling of carbonate system properties and could feed directly into
 1052 DA applications.

1053 There is an important subtlety to the application of data-assimilative models when quantifying
 1054 net CO₂ uptake due to OAE, which is highly relevant for MRV. When the net CO₂ uptake is
 1055 quantified by calculating the difference between two simulations, one with and one without
 1056 OAE (one of these is realistic, the other counterfactual), it is not appropriate to assimilate
 1057 biogeochemical observations of properties affected by the alkalinity enhancement. The
 1058 assimilation of alkalinity-related observations to constrain one of the simulations in the pair
 1059 would eliminate the ability to make comparisons between the two. However, assimilation of
 1060 observations that are unaffected by OAE (e.g., temperature, salinity, oxygen, etc.) can be
 1061 applied to both simulations of the pair. Further research and method development are required
 1062 to identify the best approaches for leverage DA in this context.

1064 3.4 Uncertainty analysis

1065 Model results should be paired with sound qualitative and quantitative uncertainty estimates,
 1066 especially when used for practical decisions. Estimating the uncertainty of model simulations,
 1067 however, is inherently difficult because typically one is most interested in simulation outputs
 1068 for which observations are not available (e.g., unobserved or insufficiently observed properties
 1069 or fluxes in the past, properties and fluxes in the future); hence, standard procedures and
 1070 metrics for model validation (Section 3.2) are not helpful for this aspect. Uncertainty estimates
 1071 could be based on extensive model parameter and configuration sensitivity studies and
 1072 comparisons with models that include more realistic representations of uncertain or

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In situ measurements of the carbonate system are much more limited temporally and especially spatially than remote sensing observations from satellites, which are the backbone of physical (satellite sea surface temperature and sea surface height) and biogeochemical data assimilation (see above). Observations of pH, pCO₂, alkalinity, and DIC used to be limited to moorings and research cruises but have more recently been extended by automated observing systems, such as gliders, BGC-Argo floats and uncrewed surface vehicles (Bushinski et al. 2019).

1090 parameterized processes. Furthermore, since specification of uncertainty is an integral part of
1091 DA, DA methodologies provide a useful framework for estimating uncertainty, especially
1092 ensemble-based methods.

1093
1094 Any DA application requires uncertainty specification of the observations that are assimilated
1095 and can provide uncertainty estimates of the results of the assimilation procedure. Specification
1096 of uncertainty in the input data is necessary to inform the DA machinery about how much
1097 weight and reach each data point or data type should have in influencing the outcome. The
1098 more realistic the uncertainties of the input data, the better the DA outcomes in terms of
1099 explanatory or predictive skill. It is important to note that “better” does not mean more precise
1100 in this context. Overconfidence in the accuracy of assimilated observations will lead to
1101 overfitting and a degradation of predictive skill. In the case of parameter optimization, the
1102 output of the assimilation exercise is a set of optimized parameters. The uncertainty of optimal
1103 parameters, referred to as *a posteriori* errors, is determined by a Hessian analysis of the cost
1104 function in combination with the uncertainty of the input parameters before optimization, the
1105 so-called *a priori* errors (Thacker et al. 1989, Fennel et al. 2001). In the case of ensemble-based
1106 state estimation, the ensemble spread of the reanalyzed model state provides a spatially and
1107 temporally resolved estimate of the uncertainty of the reanalysis (Yu et al. 2018, Hu et al. 2012).

1108
1109 However, an important caveat is that subjectivity enters the uncertainty specification in all of
1110 these approaches. For example, in the case of parameter optimization the assumed *a priori*
1111 errors, their probability distributions, and the choice of the cost function are subjective and
1112 influence the *a posteriori* errors (but interestingly the values of the observations themselves do
1113 not). In the case of ensemble-based state estimation, the sources of uncertainty inherent in the
1114 model simulation have to be specified and simulated by generating variations within a model
1115 ensemble. Sources of uncertainty include errors in atmospheric forcing and boundary
1116 conditions, model parameters, and structural uncertainty. Uncertainty in forcing and boundary
1117 conditions is often represented by perturbing the time of sampling, uncertainty in parameters is
1118 represented by sampling from a probability distribution (based on *a priori* assumptions about
1119 the uncertainty of each parameter), and the structural uncertainty is typically represented via
1120 brute-force inflation factors that amplify ensemble spread. Yu et al. (2019), Li et al. (2016), and
1121 Thacker et al. (2012) provide examples where different sources of model uncertainty are
1122 accounted for. While the mechanics by which the model ensemble is generated and spreads
1123 over time is thus subjective, grossly inappropriate choices will lead to obviously wrong or
1124 degraded reanalyses. The success of a DA exercise, which is best judged by an evaluation of
1125 whether the predictive power of the model has improved, thus provides a useful reality check
1126 on whether the choices for specifying uncertainty were appropriate.

1127
1128 How can the framework for specifying and estimating uncertainty from model ensembles be
1129 applied in the context of OAE research? Two different cases should be considered here: 1)
1130 model applications where the absolute value of quantities matters for the research question to
1131 be addressed and thus the uncertainty of the simulated output, and 2) applications where
1132 information about the difference between a simulation with and without OAE is of interest and

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the uncertainty of this difference (e.g., the net CO₂ uptake and its uncertainty in the context of MRV). Examples of the first case include studies of the stability of added alkalinity (i.e., simulation of runaway calcium carbonate precipitation) and studies about the exposure of planktonic and benthic communities to high pH. In this case, the ensemble framework described above can be applied with the caveat that the specification of all the relevant sources of uncertainty is by no means trivial and subjective to some degree.

The second case is highly relevant for MRV of OAE where one is interested in accurately quantifying the increase in seawater DIC due to OAE with well characterized uncertainty. In this case, one would use two simulations that are based on an identical model set-up with only one difference, namely a source of alkalinity is applied to one (i.e., one of these two simulations is counterfactual or hypothetical, the other would typically be as realistic as possible). It may be tempting, and is conceptually straightforward, to apply the ensemble framework for each model of the pair and combine the resulting uncertainties via error propagation. However, in practice this would not provide meaningful estimates because there are sources of uncertainty that are unaffected by OAE (e.g., atmospheric forcing) and accounting for them may significantly overestimate uncertainty in the estimated net CO₂ uptake. A more appropriate approach would be to construct an ensemble of model pairs that explicitly accounts for uncertainty related to the impacts of alkalinity addition. How to specify and simulate the sources of uncertainty directly resulting from OAE in practice remains an open research question.

4 Model experimentation

In this section, we lay out general objectives for model experimentation in the context of OAE research and provide a short historical view of how these model studies have evolved (Section 4.1) followed by specific recommendations for Observing System Simulation Experiments (Section 4.2) and model intercomparisons (Section 4.3).

4.1 General objectives of model experimentation

General objectives of OAE modeling include (1) gaining a better understanding of the biogeochemistry of OAE, including its effectiveness and side effects, (2) supporting experiments, field trials, or commercial deployments including through the optimization of observing systems, (3) assessing global carbon-cycle and climate feedbacks, (4) understanding the role that OAE can play in climate mitigation efforts, and (5) supporting monitoring, reporting, and verification activities. At a conceptual level, model approaches for OAE can be classified as belonging into one of two groups: idealized or realistic. Idealized modeling approaches are typically driven by research questions of a fundamental nature and aim to develop or test hypotheses or provide improved process understanding while strongly simplifying a range of potentially complicating factors. They are useful for illustrating cause-and-effect relationships and the range of plausible outcomes given strong assumptions. In contrast, realistic modeling approaches aim to include a broad range of contributing factors as accurately as possible and provide detailed hindcasts or predictions that, if the model has skill,

1176 can be used for a range of practical applications. In practice, the dividing line between idealized
1177 and realistic models is blurry. Of course, no model will ever simulate all aspects of reality,
1178 hence even realistic simulations make many assumptions and are prone to errors from multiple
1179 sources. It can be effective to apply idealized and realistic approaches in a complementary
1180 manner and iteratively.

1181 It is illustrative to review briefly how modeling for OAE research has developed over the course
1182 of the last decade. Much of the early work on OAE used idealized models. Model simulations
1183 were designed to investigate whether the theoretical concept of OAE could remove large
1184 amounts of CO₂ on the global scale. Rather than trying to account for the technical and socio-
1185 economic constraints of OAE deployment, the model experiments were designed to investigate
1186 what would happen if surface alkalinity was homogeneously increased by massive amounts via
1187 a constant addition rate over extremely large regions of the ocean, e.g., in all sea-ice free waters
1188 (Paquay and Zeebe, 2013; Keller et al., 2014; Ilyina et al., 2013; Köhler et al., 2010; Köhler et al.,
1189 2013). These simulated OAE deployments will never be realized, but the model results
1190 suggested that OAE can be viable as a CDR approach. A particular advantage of this idealized
1191 approach is that the effect of OAE was easy to detect against internal model variability, i.e., the
1192 signal to noise ratio is high. The next steps in modeling OAE have remained idealized but have
1193 begun to introduce more constraints and better mechanistic or empirically derived components
1194 as experimental OAE data becomes available. Recently, modeling studies tailored to specific
1195 regions and modes of application have been conducted to support field trials or commercial
1196 deployment (Mongin et al. 2021, Wang et al. 2023). These applications must be as realistic as
1197 possible. None of the modeling studies published to date have simulated an actual OAE field
1198 trial.

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1199 4.2 Recommendations for Observing System Simulation Experiments (OSSEs)

1200 Observing system simulation experiments (OSSEs) use data-assimilative simulations to design
1201 new, or modify existing, observing systems such that deployments of observing assets, e.g.,
1202 floats, gliders, moorings, or surface vehicles, is optimized. General overviews and best practices
1203 for OSSEs are provided by Halliwell et al. (2015) and Hoffman and Atlas (2016). Examples of
1204 applications to biogeochemical models include Ford (2021), Wang et al. (2020), and Denvil-
1205 Sommer et al. (2021). Their goal is to maximize the information gained from a new or modified
1206 observing system, while keeping the number of required instruments, sensors, or deployments
1207 – and thereby cost and effort – low. OSSEs are especially valuable tools in the context of OAE
1208 research because the marine carbonate system is still undersampled, observing systems need to
1209 be designed and expanded, and new instruments deployed and configured (Boyd et al. 2023).

1210 In practice, this is done with the help of a pair of two different models or model versions, also
1211 referred to as twin experiments, as follows. A simulation of one of the models is considered to
1212 be the “truth.” This simulation is also referred to as the “nature run” and synthetic observations
1213 are generated by subsampling this nature run. This subsampling can be repeated with different
1214 sampling schemes (e.g., different variable types, different numbers of profiles, transects, and/or
1215 fixed location time series, etc.) to represent different configurations of the observing system.

1217 Finally, the synthetic observations are assimilated into the other model for which a non-
1218 assimilative simulation, the so-called “free run,” is also available. The skill of this data-
1219 assimilative simulation, also referred to as the “forecast run,” can be assessed against the free
1220 run using independent observations that are also sampled from the nature run. In this way the
1221 impact of different sets of observations on the data-assimilative model can be measured and
1222 assessed.

1223 While conceptually straightforward, care and consideration are required when setting up
1224 OSSEs. For example, the choice of the two model versions making up the twin is important. If
1225 the models chosen for the truth and forecast runs are versions of the same model
1226 implementation that were generated by perturbing initial, forcing or boundary conditions in
1227 one of them, the method is referred to as the “identical twin” approach. If two different model
1228 types are used, they are “non-identical twins.” The intermediate approach where the same
1229 model type is used but in different configurations (e.g., different physical parameterizations
1230 and/or spatial resolution) is referred to as fraternal twin. The identical twin approach has been
1231 more common in oceanic DA applications although atmospheric OSSEs have shown that it can
1232 provide biased impact assessments (Hoffman and Atlas, 2016) typically because the error
1233 growth rate between the truth and forecast runs is insufficient. A direct comparison of the non-
1234 identical and identical twin approach for an ocean circulation model of the Gulf of Mexico has
1235 been conducted by Yu et al. (2019). In their assessment of the impacts of the existing observing
1236 system (consisting of satellites and Argo floats), the identical twin approach provided overly
1237 optimistic improvements in model skill after assimilation of data from some observing assets
1238 (specifically sea-surface height and temperature) but undervalued the contribution from
1239 temperature and salinity profiles. They concluded that skill assessments and OSSEs using the
1240 non-identical twin approach are more robust. Similar concerns likely apply to OSSEs for
1241 biogeochemical properties, but this remains to be studied systematically.

1242 **4.3 Recommendations for intercomparisons**

1243 A common approach to assessing model uncertainty are coordinated, multi-model studies,
1244 commonly called model intercomparison projects or MIPs. They can be used to explore the
1245 simulated range of model behaviors, to isolate the strengths and weaknesses of different models
1246 in a controlled setting, and to interpret, through idealized experiments, inter-model differences
1247 (IPCC 2013). Carefully designed experiments can also offer a way to distinguish between errors
1248 particular to an individual model and those that might be more universal and should become
1249 priority targets for model improvement (IPCC 2013). These studies rely on common agreed-
1250 upon protocols for simulating certain processes and writing of diagnostic output to ensure that
1251 best practices are followed, and results are comparable (e.g., Griffies et al., 2016). The best-
1252 known model intercomparison project is probably the Coupled Model Intercomparison Project
1253 (CMIP, Eyring et al., 2016), which is currently finishing up its 6th phase. Within CMIP6, the
1254 carbon dioxide removal intercomparison project (CDRMIP; Keller et al., 2018) is the first project
1255 to develop a model intercomparison experiment for ocean alkalinity enhancement. This and
1256 other MIP examples, including those conducted at smaller region scales (Wilcox et al., 2022),
1257 provide a blueprint for developing coordinated multi-model experiments.

1258 The following key practices have proven useful in previous coordinated multi-model
 1259 comparisons. Since broad participation is typically desired, the protocol should be
 1260 straightforward for modeling groups to implement, otherwise few will have the resources to
 1261 participate. In practice this means avoiding new implementations of complex code or requiring
 1262 too many or too long simulations. If applicable, forcing data should be centrally prepared and
 1263 provided to participants in a standardized way that enables easy modification or reformatting,
 1264 if needed, for use with different models. Using common simulations that modeling groups are
 1265 likely to have completed already, e.g., climate change scenarios, as control runs and
 1266 experimental branching points is helpful for minimizing the number of additional required
 1267 simulations. It is useful to establish common practices that facilitate the production and analysis
 1268 of the model output, e.g., what should be archived and shared (Jukes et al., 2020) and data
 1269 standards governing the structure and required metadata for model output (Pascoe et al., 2020).
 1270 Shared software to standardize model output, such as the Climate Model Output Rewriter
 1271 (CMOR; <https://cmor.llnl.gov/>) commonly used in CMIP, can be helpful. To maximize the use of
 1272 model output, it should be made available for public download with digital object identifiers
 1273 (DOIs). The Earth System Grid Federation (ESGF) is an example of such a system (Petrie et al.,
 1274 2021). If applicable, preparing and providing quality-controlled observational datasets for
 1275 model evaluation is useful for facilitating analytical efforts (Waliser et al., 2020). Coordinating
 1276 the analysis is helpful to avoid duplicative efforts and ensure consistent application of
 1277 evaluation metrics. Finally, the design of a coordinated multi-model experiment and all its
 1278 procedures should be well documented in publications or permanently archived protocols. It is
 1279 advisable to test the multi-model experiment with a small subset of models, before inviting a
 1280 large number of participants. Furthermore, it is worth remembering that the science questions
 1281 must be appropriate. MIPs require much effort and not every science question needs a MIP to
 1282 be answered.

1283 5 Summary and Key Recommendations

1284 A range of modeling tools and analysis methods are available for OAE research to address
 1285 questions from micro- to global scales; however, each of these tools and methods has limitations
 1286 and caveats that model users and users of model-generated outputs need to be aware of.
 1287 Furthermore, this new field of research poses questions and challenges that current tools were
 1288 not designed to address, necessitating further development.

1289 A common objective of all modeling approaches described in this [article](#) is to simulate the
 1290 spatio-temporal evolution of carbon chemistry properties in seawater by accounting for the
 1291 physical, chemical, and biological processes that determine this evolution. Idealized models,
 1292 which neglect some aspects of reality in the interest of simplicity and clarity of assumptions,
 1293 have long been used to test basic questions about OAE. As research questions are becoming
 1294 more focussed on the practical aspects, feasibility, and ecosystem impacts of OAE, more realistic
 1295 models are increasingly desirable. A skillful realistic model can provide spatial and temporal
 1296 context for observations, including estimates of properties and fluxes not directly observed.
 1297 Such model will include parameterizations of the relevant processes for the research objective to
 1298 be addressed and will be constrained by observations that contain sufficient meaningful

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1314 information. However, model formulations of several properties and processes relevant to OAE
1315 research remain uncertain or highly simplified. For example, presently used model
1316 representations of alkalinity in seawater are likely inadequate and may require explicit
1317 representation of at least some of the multiple biotic and abiotic sources and sinks of alkalinity;
1318 the mechanisms and triggers for spontaneous calcium carbonate precipitation are only
1319 beginning to be described and not yet represented in models; and the impacts of pH
1320 perturbations on plankton diversity and trophic interactions remain an active area of study and
1321 unaccounted in biogeochemical models. Furthermore, it is difficult to obtain solid constraints on
1322 the seawater carbonate system, especially in sufficient spatial and temporal resolution for
1323 robust model validation and DA. Theoretically, knowledge of two of the carbonate system
1324 parameters allows calculation of the others, but unfortunately $p\text{CO}_2$ and pH, the pair most
1325 accessible to autonomous measurement, results in high uncertainties.

1326 One inherent challenge to OAE research is the multiscale nature of many of the relevant
1327 questions. Different modelling tools are available for different spatial scales. While some
1328 research questions may fall neatly within the limited spatial range of a particular model, many
1329 do not and require a bridging of scales that could be accomplished via new parameterizations
1330 yet to be developed or dynamic coupling of different modeling tools. It is important to
1331 emphasize that models have to be tailored to the questions they are meant to address. This
1332 means considering what level of model complexity is required and seeking parsimonious
1333 representations that are well-supported by empirical constraints.

1334 It is important to note that even after thorough validation, any model simulation will be prone
1335 to errors and uncertainties due to inaccuracies in model inputs, structural uncertainty due to
1336 numerical approximation schemes and insufficient process understanding or representation,
1337 and inaccurate model parameters and parameterizations. Deviations between models and
1338 reality can be reduced by DA, which is typically applied either to systematically optimize
1339 model parameters or to produce optimal estimates of the ocean state. Optimization of model
1340 parameters addresses systematic model errors and biases; it is useful for systematic testing of
1341 different model formulations during model design. State estimation assumes an unbiased
1342 model and addresses unresolved stochasticity, thus leading to model states that are in better
1343 agreement with the observed ocean state. However, successful application of DA critically
1344 requires sufficient observations. Currently, the biggest impediment to implementing data-
1345 assimilative models for OAE research is the sparsity of carbonate system observations. OSSEs,
1346 data-assimilative simulations that inform how to place observing assets most effectively, will
1347 prove useful in this context. It should also be noted that assimilation of carbonate system
1348 parameters is not appropriate when models are applied for MRV.

1350 Uncertainty analysis is a necessary component of any quantitative research and will be an
1351 essential deliverable for effective approaches to MRV. Ensemble-based DA methodologies
1352 provide a useful framework for estimating uncertainty. Consideration of this framework
1353 illustrates the “law of conservation of difficulty” applies here. Quantitative assumptions about
1354 the uncertainty distributions of input data and input parameters, and of structural uncertainties
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inherent in the model are required to obtain an uncertainty estimate of the model output, in other words, difficult assumptions about errors have to be made somewhere. A common approach to assessing model uncertainty is by coordinated, multi-model intercomparison. Such studies can be used to explore the range of simulated behaviors and the strengths and weaknesses of different models and, by elucidating inter-model differences, they can offer guidance on priority targets for model improvement.

Key recommendations arising from this article are as follows:

- Idealized models of particle-fluid interaction are recommended to address questions about dissolution and precipitation kinetics at the scale of particles, realistic local-scale models are recommended for addressing questions about nearfield processes in the turbulent environment around injection sites, and larger-scale regional or global ocean models are recommended to support observational design for field experiments, to demonstrate possible verification frameworks, and to address questions about global-scale feedbacks on ocean biogeochemistry.
- When simulating OAE approaches that may generate high oversaturation with respect to carbonate, spontaneous precipitation of carbonates needs to be considered and appropriate approaches should be developed, e.g., using near-field models to mechanistically represent this process and a meta-model approach to develop parameterizations that are suitable for far-field and larger-scale models.
- Shortcomings in current-generation models in terms of representing physiological responses of the plankton community to OAE (especially when using crushed-rock feedstocks) need to be recognized, better qualified, and addressed. Empirical research exploring physiological sensitivities should be used to develop prioritizations of key model processes comprising early targets for implementation.
- The exchange of solutes between the sediments and overlying water influences the seawater carbonate system with DIC from the remineralization of organic matter being returned to overlying water (and alkalinity if this remineralization occurs anaerobically), dissolution of CaCO_3 releasing alkalinity, and burial of CaCO_3 acting as alkalinity sink. Accounting for these exchanges between sediments and overlying water and its variability on tidal, seasonal, interannual, and millennial timescales will likely be necessary in regional and global biogeochemical models that aim to simulate alkalinity cycling.
- River inputs of alkalinity and DIC in regional and global ocean biogeochemical models, including fluxes resulting from land-based CDR applications, should be accurately accounted for. Efforts should be made to improve quantification of riverine fluxes resulting from ongoing field trials and commercial applications, and to establish initiatives to effectively track the solute additions from terrestrial alkalinity enhancements.

- When simulating large-scale deployment of OAE in ocean-only models with prescribed atmospheric CO₂, the subtle changes in the atmospheric CO₂ inventory resulting from CDR should be accounted for.
- Models should be tailored to the specific questions they are meant to address while seeking parsimonious representations that are well-supported by empirical constraints. For example, for the purpose of MRV it may be appropriate to neglect biological dynamics since the core target is to capture the net air-sea CO₂ exchange associated with the OAE-induced surface ocean pCO₂ anomaly.
- Model validation should be an integral part of model implementation and application. For OAE research, validation is a two-step challenge. First, it is necessary to validate unperturbed model baselines to gain confidence that the natural variability is represented appropriately and to quantify model uncertainties. Second, it should be verified that the model accurately represents the perturbations of an alkalinity addition.
- Since no single model validation metric provides a complete picture of a model's skill, multiple complementary metrics should be used in combination. Furthermore, different points in space and time, and a breadth of variable types should be part of any comprehensive validation.
- Data assimilation, the process of improving the dynamical behavior of models by statistically combining them with observations, should be employed in order to obtain the most accurate model simulations possible, e.g., to optimize model parameters or to estimate the ocean state. The former addresses systematic errors and biases in models, while the latter assumes an unbiased model and addresses unresolved stochasticity.
- When applying data-assimilative models for quantification of the OAE-induced net CO₂ uptake by calculating the difference between a realistic and a counterfactual simulation, it is not appropriate to assimilate biogeochemical observations of properties affected by the alkalinity enhancement as this would eliminate the ability to make valid comparisons between the two simulation. However, assimilation of observations that are unaffected by OAE can be applied to both simulations of the pair.
- Successful application of DA critically requires sufficient observations either of the properties that the model parameters to be estimated depend on or of the state variables that are being estimated. Observing System Simulation Experiments are recommended to design observing strategies tailored to the needs of specific OAE applications.
- Model results should be paired with sound qualitative and quantitative uncertainty estimates, especially when used for practical decisions. DA methodologies provide a useful framework for estimating uncertainty, especially ensemble-based methods. Another common approach to assessing model uncertainty are coordinated, multi-model studies, commonly called model intercomparison projects or MIPs.

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Competing interests

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