

1 **Modeling considerations for research on Ocean Alkalinity Enhancement**
2 **(OAE)**

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18

19 **Abstract**

20 The deliberate increase of ocean alkalinity (referred to as Ocean Alkalinity Enhancement or
21 OAE) has been proposed as a method for removing CO₂ from the atmosphere. Before OAE can
22 be implemented safely, efficiently, and at scale several research questions have to be addressed
23 including: 1) which alkaline feedstocks are best suited and in what doses can they be added
24 safely, 2) how can net carbon uptake be measured and verified, and 3) what are the potential
25 ecosystem impacts. These research questions cannot be addressed by direct observation alone
26 but will require skillful and fit-for-purpose models. This [article](#) provides an overview of the
27 most relevant modeling tools, including turbulence-, regional- and global-scale biogeochemical
28 models, and techniques including approaches for model validation, data assimilation, and
29 uncertainty estimation. Typical biogeochemical model assumptions and their limitations are
30 discussed in the context of OAE research, which leads to an identification of further
31 development needs to make models more applicable to OAE research questions. A description
32 of typical steps in model validation is followed by proposed minimum criteria for what
33 constitutes a model that is fit for its intended purpose. After providing an overview of
34 approaches for sound integration of models and observations via data assimilation, the
35 application of Observing System Simulation Experiments (OSSEs) for observing system design
36 is described within the context of OAE research. Criteria for model validation and
37 intercomparison studies are presented. The article concludes with a summary of
38 recommendations and potential pitfalls to be avoided.

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41 **1 Introduction**

42 Ocean Alkalinity Enhancement (OAE) refers to the deliberate increase of ocean alkalinity, which
43 can be realized either by removing acidic substances from or adding alkaline substances to
44 seawater. OAE is receiving increasing attention as a method for removing CO₂ from the
45 atmosphere; such methods are referred to as marine Carbon Dioxide Removal (mCDR)
46 technologies (Renforth and Henderson, 2017). Natural analogues to OAE exist (Shubas et al.
47 2023). An increase in the alkalinity of seawater leads to a repartitioning of its dissolved
48 carbonate species with a shift toward bicarbonate and carbonate ions (Zeebe and Wolf-Gladrow
49 2001, Renforth and Henderson 2017), leading to a reduction in the aqueous CO₂ concentration
50 and thus the partial pressure of CO₂ (pCO₂; Schulz et al. 2023). Since exchange of CO₂ between
51 the ocean and atmosphere occurs when the surface ocean pCO₂ is out of equilibrium with that of
52 the atmosphere, a lowering of the ocean's pCO₂ will lead to a net ingassing of atmospheric CO₂
53 (i.e., an increase in CO₂ uptake by the ocean or a decrease in outgassing due to OAE). This
54 would increase the oceanic and decrease the atmospheric inventories of inorganic carbon, in
55 other words, it would result in mCDR. In contrast to other mCDR technologies, OAE does not
56 exacerbate ocean acidification (Ilyina et al. 2013). In fact, an increase in ocean alkalinity
57 counteracts acidification, and while subsequent net uptake of atmospheric CO₂ largely restores
58 pH to its pre-perturbation value, there is potential for OAE deployment to mitigate acidification
59 impacts near injection sites (Mongin et al. 2021).

60 Several important research questions should be addressed before implementing OAE as an
61 mCDR technology at scale. These include: 1) which alkaline substances are best suited and in
62 what doses can they be added reliably while avoiding precipitation of calcium carbonate (which
63 would decrease alkalinity and could result in runaway precipitation events), 2) how can
64 changes in alkalinity and net carbon uptake be measured, verified, and reported (referred to as
65 MRV; see Ho et al. 2023) to enable meaningful carbon crediting, and 3) what are the potential
66 ecosystem impacts and how can harm to ecosystems be avoided or minimized while
67 maximizing potential benefits. These research questions cannot be addressed by direct
68 observation alone, but will require an integration of observations and numerical ocean models
69 across a range of scales. Skillful and fit-for-purpose models will be essential for addressing
70 many OAE research questions including the MRV challenge, assessment of environmental
71 impacts, and interpretation of natural analogs.

72 Ocean models are useful for a broad range of purposes, from idealized models for basic
73 hypothesis testing of fundamental principles to realistic models for more applied uses (see
74 primer on ocean biogeochemical models by Fennel et al. 2022). In the context of OAE research,
75 this full range of models is applicable. For example, idealized models of particle-fluid
76 interaction can inform us about dissolution and precipitation kinetics at the scale of particles,
77 realistic local-scale models can inform us about nearfield processes in the turbulent
78 environment around injection sites, and larger-scale regional or global ocean models can be
79 used to support observational design for field experiments, to demonstrate possible verification
80 frameworks, and to address questions about global-scale feedbacks on ocean biogeochemistry.
81 A common objective of all these modeling approaches is to realistically simulate the spatio-

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88 temporal evolution of the seawater carbon chemistry, including alkalinity and dissolved CO₂,
89 and attribute that evolution to physical, chemical, and biological processes. Models that are
90 suitable for this purpose will provide spatial and temporal context for properties that can be
91 observed (but at much sparser temporal and spatial coverage than a model can provide) as well
92 as estimates of properties and fluxes that cannot be directly observed but may be inferred
93 because of known mechanistic relationships or patterns of correlation. Applications of realistic
94 models rely on them being skillful and accurate, requiring that they include parameterizations
95 of the relevant processes, and that they are constrained by observations that contain sufficient
96 meaningful information (what is sufficient depends on the application and research question).
97 Methods for constraining models by observations through statistically optimal combination of
98 both are available. Application of such methods is referred to as data assimilation and provides
99 the most accurate estimates of biogeochemical properties and fluxes (see Fennel et al. 2022 for
100 fundamentals and code examples).

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101 Model applications for OAE research include the following four general types:

- 102 • Hindcasts are model applications where a defined time period in the past was
103 simulated. They can be unconstrained—in the sense that no observations are fed into the
104 model except for initial, boundary, and forcing conditions—or constrained, where
105 observations inform the model state via data assimilation. The latter are also referred to
106 as optimal hindcasts or reanalyses.
- 107 • Nowcasts/forecasts are similar to constrained hindcasts but with the simulations carried
108 out up to the present (referred to as nowcasts) or into the future (referred to as
109 forecasts). The latter require assumptions about future forcing and boundary conditions,
110 e.g., from other forecasts, climatology, or assuming persistence.
- 111 • Scenarios are unconstrained hindcasts or forecasts where one or more aspects of the
112 model is systematically perturbed to assess the effect of the perturbation, for example, in
113 paired simulations with and without OAE, one would be the realistic case and the other,
114 a scenario (also referred to as counterfactual in this case). These can be used to explore
115 even very unlikely situations, which is often required in comprehensive uncertainty and
116 risk assessment.
- 117 • Observing System Simulation Experiments (OSSEs) for observing system design use
118 unconstrained and/or constrained hindcasts to evaluate the benefits of different
119 sampling designs and optimize deployment of observational assets for a defined
120 objective, including tradeoffs between different types of observation platforms.

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121 Successful implementation of models to support OAE research and MRV is challenging because
122 of the general sparseness of relevant biogeochemical observations, and the limited lab,
123 mesocosm, and field trial data available to date for model parameterization. Further, models are
124 built at a process level and integrated to reveal behavior at the emergent scale. As such, models
125 comprise a collective hypothesis of the ocean's physical, biogeochemical, and ecosystem
126 function—but it is important to recognize that model formulations of key processes related to

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135 OAE remain uncertain. It may well turn out that parameterizations of the carbonate system, of
136 plankton diversity and trophic interactions, small scale turbulence, submesoscale subduction
137 and restratification processes, and air-sea gas exchange in the current generation of models
138 require improvement to robustly treat OAE-related questions.

139 The intended scope of this [article](#) is to provide an overview of the most relevant modeling tools
140 for OAE research with high-level background information, illustrative examples, and references
141 to more in-depth methodological descriptions and further examples. We aim to provide simple
142 criteria and guidance for researchers on the current state-of-the-art of biogeochemical modeling
143 relevant to OAE research, keeping in mind short-term research goals in support of pilot
144 deployments of OAE and long-term goals such as credible MRV in an ocean affected by large-
145 scale deployment of OAE and possibly other CDR technologies.

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146 2 Modeling approaches

147 This section provides a brief review of modeling tools available for OAE research with
148 references to more in-depth methodological descriptions and examples, as well as a discussion
149 of which approaches are most applicable to simulating essential processes in different
150 circumstances. The presentation is structured using two complementary organizing principles,
151 the spatial and temporal scales of the problem in Section 2.1 and the biogeochemical and
152 ecological complexity represented by different modeling approaches in Section 2.2. [Section 2](#),
153 concludes with a summary of suggested future model development efforts in Section 2.3.
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155 2.1. Modeling approaches across scales

156 In the nearfield, close to the site of an alkalinity increase, an accurate characterization of the
157 spatio-temporal evolution of alkalized waters requires direct representation or parameterization
158 of fluid and particle physics and seawater carbonate chemistry at scales ranging from
159 micrometers to hundreds of meters, spanning turbulent to submesoscale processes (Section
160 2.1.1). In the farfield, covering scales from 10s of meters to 100s of kilometers, where the effect of
161 an alkalinity increase depends less on the details of how the alkalinity was added, or acidity
162 removed, and is instead dominated by ambient environmental processes, local to regional scale
163 models are useful for simulating the impact of alkalinity increases, for verifying the intended
164 perturbations in air-sea exchange of CO₂ and in carbonate system variables, and potentially for
165 simulating ecosystem impacts (Section 2.1.2). Lastly, investigation of the effects of the global
166 ocean's overturning circulation, impacts on atmospheric CO₂ levels, and of Earth system
167 feedbacks resulting from deployment of OAE and other CDR technology at scale requires
168 global modeling approaches (Section 2.1.3).

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169 2.1.1. Particle scale to nearfield/turbulence scale (µm to km scales)

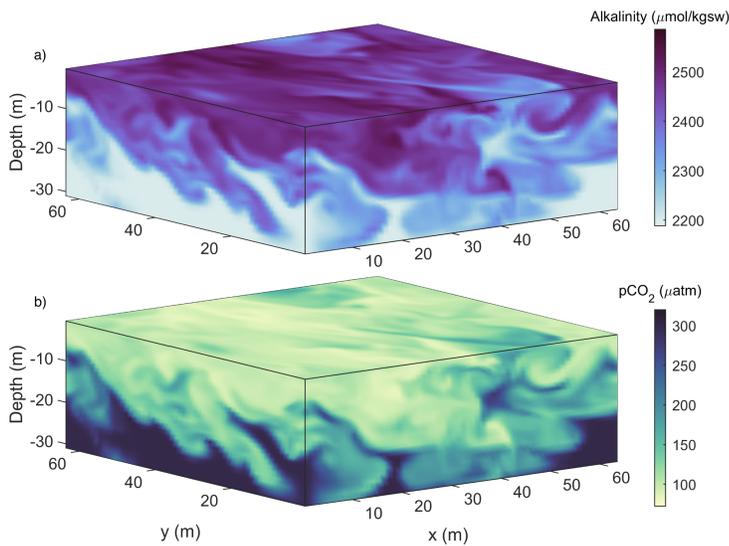
171 Small-scale modeling approaches cover the range from µm-size particles to the turbulent- and
172 submeso-scales in the nearfield of alkalinity additions. Simulating processes on these scales
173 allows one to address questions about how turbulent mixing dilutes and disperses alkalized
174 water and how it affects the settling, aggregation, disaggregation, precipitation, and dissolution

179 of suspended particles. Nearfield modeling has an important role to play in guiding the design
180 of deployment strategies that mitigate environmental impacts and meet future permitting
181 requirements, and to support monitoring. During the initial dispersion and dilution phase of an
182 alkalinity increase in the nearfield, the direct impacts on carbonate system variables are
183 greatest, with waters exhibiting the largest elevations in pH and the highest potential for the
184 formation of secondary precipitates. For particulate alkalinity feedstocks, turbulence close to the
185 deployment site affects dissolution and settling rates, increasing dissolution and either
186 accelerating or diminishing the settling of sedimentary particles compared to the Stokes settling
187 speed (Fornari et al. 2016).

188 Distinct approaches to modeling at these scales involve different levels of parametrization and
189 computational expense, with the relative utility of each approach being dependent on the
190 scientific questions at hand. At the smallest scales, Direct Numerical Simulations (DNS) are the
191 most computationally expensive and specialized class of fluid modeling, as they resolve flows
192 down to the scales at which flow variances dissipate—typically centimeters or smaller in the
193 ocean. Consequently, computational constraints imply that they cannot be run over domains
194 larger than a few meters. DNS are thus integrated over idealized physical domains (i.e., they
195 lack realistic bathymetry) and are suited to investigating fundamental physical processes. For
196 example, multiphase DNS simulations have been used to model the interaction of turbulence
197 with gas bubbles (Farsoiya et al. 2023) and particles (Fornari et al. 2016). Results from such
198 studies provide an important testbed that can be used to develop parameterizations required in
199 lower resolution models.

200 A well-established approach to modeling the fluid flow at scales up to about 10 km uses Large
201 Eddy Simulations (LES), a class of model that directly solves the unsteady Navier-Stokes
202 equations down to the largest turbulent scales on a high-resolution grid. Such models
203 parameterize turbulence using a subgrid-scale model (e.g., Smagorinsky 1963). An advantage of
204 these models is their ability to simulate both an alkalized plume and the environmental
205 turbulence into which the plume emerges. Once alkalized waters enter the surface boundary
206 layer, LES models have an established history of simulating turbulence and mixing that is
207 directly relevant to OAE research (e.g., Mensa et al. 2015, Taylor et al. 2020). An example of an
208 LES simulation of near-surface turbulence dispersing surface-deployed alkalinity downwards is
209 illustrated in Figure 1, where a physical model (Ramadhan et al. 2020) has been coupled to a
210 carbonate solver (Lewis et al. 1998). To date, LES models have rarely been coupled to
211 biogeochemical models due to the computational expenses involved, though their inclusion
212 may be increasingly feasible (Smith et al. 2018, Whitt et al. 2019). As LES simulate flow physics
213 at scales ranging from 10-10,000 m, they do not explicitly resolve the microscales of fluid motion
214 and chemical reactions at particle scales. Nevertheless, the parameterizations of such processes
215 can be included; for example, Liang et al. (2011) used models of bubble concentration and
216 dissolved gas concentration in an LES to examine the influence of bubbles on air-sea gas
217 exchange.

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 220 **Figure 1:** LES of near surface turbulence coupled to a carbonate system solver. Alkalinity is
 221 added at a rate of $4 \mu\text{mol kgsw}^{-1} \text{m}^{-2} \text{s}^{-1}$ for 20 minutes to the top grid cell at the start of the
 222 simulation. Turbulence, generated by surface wind stress and cooling, sets the rate at which it
 223 mixes downwards (a) along with associated waters of lowered $p\text{CO}_2$ (b). Turbulent plumes and
 224 eddies lead to inhomogeneities in water properties at scales of tens of meters.

225
 226 For alkalized plumes associated with outfalls from, for example, wastewater treatment plants,
 227 integral models (that assume plume properties such that the governing equations are
 228 simplified) have been developed to examine the initial dilution close to jets and buoyant plumes
 229 up to kilometer scales (Jirka et al. 1996). These models are highly configurable, enabling specific
 230 diffuser configurations as well as the potential to incorporate sediment laden plumes with
 231 particle settling (Bleninger & Jirka 2004). Results are commonly accepted for engineering
 232 purposes, defining mixing zones, and providing a fast “first look” at diffusion and mixing near
 233 an outfall site. However, these models rely on assumptions about the underlying physics of
 234 fluid flow (e.g., axisymmetric plumes and simplified entrainment rates) that may not be
 235 accurate under general oceanic conditions, and results will not include all effects of irregular
 236 bathymetry, finite domain size or arbitrarily non-uniform ambient conditions. Nevertheless,
 237 their simplicity makes them very useful. For example, by combining several simple process
 238 models for plume dilution, particle dissolution, and carbon chemistry, Caserini et al. (2021)
 239 have simulated the initial dilution of slaked lime $\text{Ca}(\text{OH})_2$ particles and alkalinity in a plume
 240 behind a moving vessel.

241 Other methods for modeling at this scale include Reynolds Averaged Navier Stokes (RANS)
 242 and Unsteady RANS (URANS), wherein fluctuations against a slowly varying or time mean

243 background are parametrized, often using constant (large) eddy diffusivities and viscosities.
244 These approaches are often inaccurate at these scales, resulting in simulations that are too
245 diffusive or lacking processes that are of leading order importance to mixing (Golshan et al.
246 2017, Chang & Scotti 2004).

247 There are multiple, potentially interacting sources of uncertainty to consider when evaluating
248 the uncertainty of the applications described above. Perhaps best understood but still
249 problematic is the uncertainty that arises from the computational intractability of simulating all
250 the relevant scales in the μm to km range at once, necessitating the different modeling
251 approaches for different scales, with parameterizations to account for unresolved scales and
252 scale interactions. The dissolved carbonate chemistry of seawater is relatively well
253 parameterized (Zeebe and Wolf-Gladrow 2001), but some modest uncertainties arise from
254 approximations required for computational tractability (Smith et al. 2018). The least understood
255 but potentially dominant source of uncertainty pertains to the representation of the microscale
256 biological, chemical, and physical dynamics of particles, which is an active area of experimental
257 and observational investigation (Subhas et al. 2022, Fuhr et al. 2022, Hartmann et al. 2023).

258 While the explicit multiphase modeling of the particles themselves is computationally costly, an
259 approach wherein the parametrized evolution of inertia-less Lagrangian particles are simulated
260 may provide a fruitful middle ground, providing a mechanism to realistically determine the
261 alkalinity release field associated with the advection, mixing, sinking and dissolution of reactive
262 mineral particles. These questions about particles apply to those released in OAE deployments,
263 as well as particles that precipitate from seawater in part due to OAE deployments, and finally
264 the role of ambient biotic and abiotic particles where OAE is deployed.
265

266 2.1.2. Local to regional scales (m to km)

267 Local to regional scale models that range in horizontal resolution from tens of meters to
268 hundreds of kilometers are useful for simulating the impact of alkalinity injections beyond the
269 immediate local area, where conditions do not depend on the details of how the alkalinity was
270 added and instead are determined by regional-scale currents and other process, including the
271 potential for biogenic feedbacks. These models are particularly useful to support OAE field
272 experiments, including planning and observational design, and analysis, integration and
273 synthesis of observations, and to facilitate interpretation of observations from natural analogs.
274 Furthermore, local and regional scale models will likely prove to be indispensable for
275 quantification of OAE effects in research settings, for guiding assessments of its environmental
276 impacts, and for MRV during the potential implementation of OAE. A skillful model can
277 simulate when and where changes in carbonate chemistry and the ensuing anomalies in air-sea
278 CO_2 exchange occur and provide an estimate of the spatio-temporal extent of the
279 biogeochemical properties affected by OAE.
280

281 Regional models have distinct advantages over global models in their ability to resolve the
282 spatial scales on which OAE would be applied both experimentally and operationally, and their
283 documented skill in representing coastal and continental shelf processes more accurately
284 (Mongin et al. 2016, Laurent et al. 2021). Examples of regional model applications in the context

285 of OAE include the recent studies by Mongin et al. (2021) and Wang et al. (2023). Mongin et al.
286 (2021) used a coupled physical-biogeochemical-sediment model tailored to Australia's Great
287 Barrier Reef to investigate to what extent realistic OAE applied along a shipping line could
288 alleviate anthropogenic ocean acidification on the reef. Wang et al. (2023) used a coupled ice-
289 circulation-biogeochemical model of the Bering Sea to study the efficiency of OAE in coastal
290 Alaska.

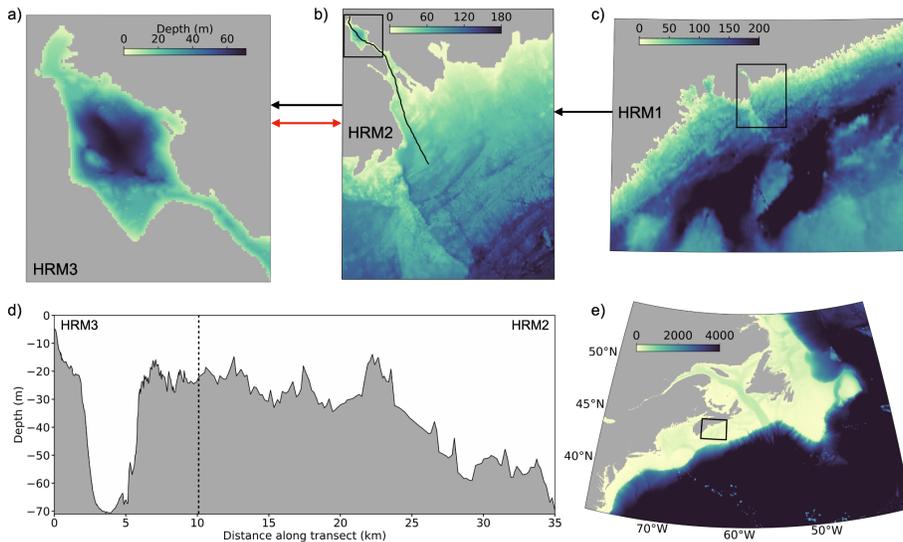
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292 Implementation of a regional model in a target domain requires generation of a grid with
293 associated bathymetry, specification of boundary conditions (including atmospheric forcing,
294 information about ocean dynamics along the lateral boundaries of the domain, any fluxes of
295 biogeochemical properties across the air-sea, sediment-water, and land-ocean boundaries, river
296 inputs), and generation of initial conditions within the domain (Fennel et al. 2022). Different
297 circulation models are available for implementation in domains targeted for OAE studies (see,
298 e.g., Table 1 in Fennel et al. 2022), all with distinct strengths and established user communities.
299 Particularly relevant in the context of studying coastal applications of OAE is a model's ability
300 to accurately represent coastal topography, making unstructured grid models and models with
301 terrain-following coordinates particularly attractive. Another feature to be considered is a
302 model's ability to run in two-way nested configurations. In the more widely applied one-way
303 nesting of domains, simulated conditions from a larger scale model (referred to as the parent
304 model) are used to generate the dynamic lateral boundary conditions of a smaller scale, higher
305 resolution model (the child model), which runs off line from the parent model. With two-way
306 nesting, both models run simultaneously and information is exchanged continually along their
307 intersecting boundaries. This allows information generated within the high-resolution child
308 domain (e.g., the spreading distribution of a tracer or alkalinity addition) to be received and
309 propagated by the larger-scale parent model. In this context, model simulations are particularly
310 useful if available in near-real time or in forecast mode. This requires specification of lateral
311 boundary conditions and atmospheric forcing up to the present and into the future. Global
312 1/12th-degree nowcasts and 10-day forecasts of ocean conditions are available from the
313 Copernicus Marine Service (CMEMS 2023) and atmospheric forcing up to the present and 10
314 days into the future are available from the European Centre for Medium Range Weather
315 Forecasts (ECMWF 2023).

316
317 One example of a high-resolution local scale model with two-way nested domains is a
318 framework developed for Bedford Basin in Halifax, Canada (Figure 2, Laurent et al., 2024). The
319 model framework consists of three nested ROMS models (ROMS is the Regional Ocean
320 Modelling System; <https://myroms.org>, Haidvogel et al. 2008, Shchepetkin and McWilliams
321 2005). The outermost ROMS domain has a resolution of 900 m and is nested one-way within the
322 data-assimilative global GLORYS reanalysis of physical and biogeochemical properties
323 (Lellouche et al. 2021). Nested within are two models with increasingly higher resolutions of
324 200 m and 60 m. Depending on the scientific objective to be addressed, the models can be run in
325 one-way and two-way nested mode, where two-way nesting is computationally more
326 demanding, and in hindcast or forecast mode. Implementation of dye-tracers within the model
327 (Wang et al. 2024) allows one to determine dynamic distribution patterns and residence times.

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333 **Figure 2:** Nested configuration of three ROMS models for the Bedford Basin and the adjacent
 334 harbor in Halifax Regional Municipality (HRM). a) The highest resolution model (HRM3; 60 m)
 335 includes the 7 km-long and 3 km-wide Bedford Basin and The Narrows, a 20-m shallow narrow
 336 channel that connects the basin to the outer harbor. b) The larger scale model (HRM2, 200 m)
 337 includes Bedford Basin and Halifax Harbor as well as the adjacent shelf. c) The largest-scale
 338 model (HRM3, 900 m) covers the central part of the Scotian Shelf as indicated in e). d)
 339 bathymetry along a section through HRM3 and HRM2, indicated by the black line in b). Lateral
 340 boundaries of HRM3, HRM2, and HRM1 are shown by black boxes in b), c) and e), respectively.
 341 Black arrows indicate the information flow between models in one way nesting mode. The red
 342 arrow indicates that HRM1 and HRM2 can be run simultaneously with bi-directional flow of
 343 information (two-way coupled mode).
 344

345 2.1.3. The global scale

346 A strength of global ocean models is their capacity to comprehensively represent the global
 347 overturning circulation and ocean ventilation. These processes control the time scales over
 348 which waters are sequestered in the ocean interior and determine how long surface waters are
 349 exposed to the atmosphere and can exchange properties, including CO₂, before being injected
 350 back into the ocean interior (Naveira Garabato et al. 2017). Similarly, the large-scale overturning
 351 circulation and the patterns associated with ventilation are important to consider in the context
 352 of deploying OAE at scale, as these patterns exert strong control on the efficiency of OAE at
 353 sequestering CO₂ (e.g., Burt et al. 2021).

354 When global ocean models are dynamically coupled with models of the land biosphere and the
355 atmosphere, they are referred to as Earth System Models (ESMs) and can be employed to
356 explore Earth system feedbacks to mCDR. In the case of OAE, the main feedback is the change
357 in atmospheric $p\text{CO}_2$ and air-sea gas exchange that will result when CDR approaches are
358 implemented at scale. While regional models have to be forced by atmospheric CO_2
359 concentrations, ESMs represent the atmospheric reservoir and are forced by CO_2 emissions into
360 the atmosphere, which then interacts with land and ocean carbon reservoirs. Only the latter
361 approach can account for OAE-induced reductions in the atmospheric CO_2 inventory which, in
362 turn, would lead to a systematic reduction in air-sea CO_2 fluxes. Regional models and global
363 ocean models that do not explicitly represent the atmospheric CO_2 reservoir and instead are
364 forced by prescribed atmospheric $p\text{CO}_2$ cannot simulate the decline in atmospheric $p\text{CO}_2$ due to
365 OAE. Depending on the alkaline material applied, there may also be feedbacks associated with
366 changes in temperature, albedo, nutrient cycles, and biological responses which can be studied
367 with the help of ESMs.

368 Another important strength of global models relates to the fact that anomalies in air-sea CO_2
369 flux generated by OAE deployments will manifest over large spatio-temporal scales because
370 CO_2 equilibrates with the atmosphere via gas exchange slowly. Alkalinity enhanced waters can
371 be transported far away from injection sites before equilibration is complete (He and Tyka
372 2023). Consequently, OAE signals may exit the finite domain of regional models prior to full
373 equilibration with the atmosphere (e.g., Wang et al. 2023). Because global models represent the
374 entire ocean and can be integrated for centuries and longer, they enable full-scale assessments.

375 A primary challenge for global models, however, is that their horizontal resolution is
376 necessarily limited by computational constraints (see example in Figure 3). Most of the global
377 ocean models contributing the Coupled Model Intercomparison Project version 6 (CMIP6), for
378 example, have horizontal resolutions of about 1° or roughly 100 km (Heuzé 2021) and do not
379 accurately represent biogeochemical processes along ocean margins (Laurent et al. 2021). Model
380 grid-spacing imposes a limit on the dynamical scales that can be explicitly resolved in the
381 models; this is particularly problematic for coarse resolution global models because mesoscale
382 eddies—i.e., motions on scales of about 10–100 km—dominate the variability in ocean flows
383 (Stammer 1997). Since coarse resolution models cannot resolve mesoscale eddies explicitly, the
384 rectified effects of these phenomena, including their role in transporting buoyancy and
385 biogeochemical tracers, must be approximated with parameterizations (e.g., Gent and
386 McWilliams 1990).

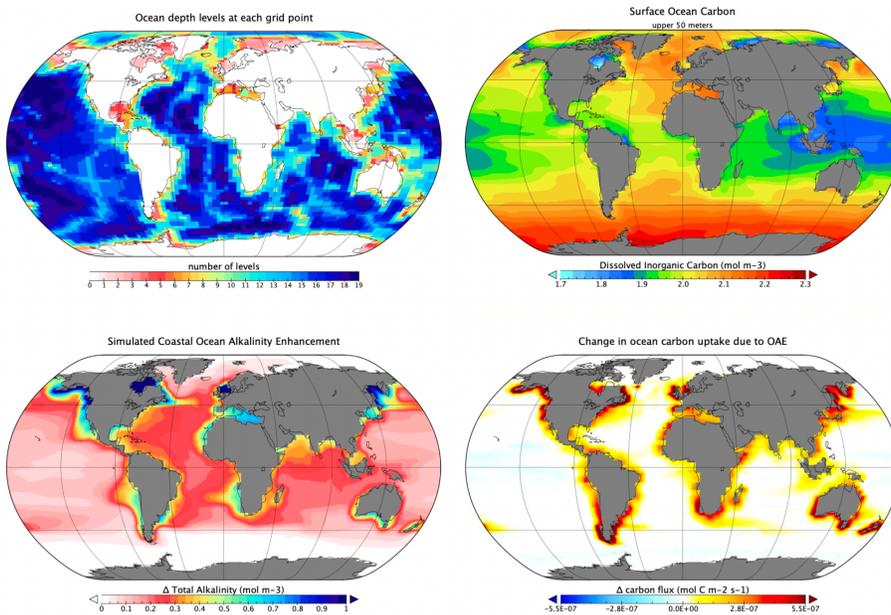
387 Notably, the fidelity of the simulated flow in global models, including the imperfect nature of
388 these parameterizations, projects strongly on the model's capacity to accurately simulate
389 ventilation and the associated uptake of transient tracers, such as anthropogenic CO_2 or
390 chlorofluorocarbons (CFCs), from the atmosphere (e.g., Long et al. 2021). Biases in the uptake of
391 transient tracers will also have implications for a model's capacity to faithfully represent the
392 impact of OAE, where the path of alkalinity-enhanced waters parcels in the surface ocean, and
393 their subsequent transport to depth is a key control on the efficiency of carbon removal. Biases
394 in the simulated flow are also an important determinant of the simulated distribution of

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399 biogeochemical tracers in the model's mean state. Hinrichs et al. (2023), for example,
400 demonstrate that inaccuracies in the physical redistribution of alkalinity by the flow is a
401 dominant mechanism contributing to biases in the alkalinity distributions simulated by CMIP6
402 models.



403

404 **Figure 3:** Example of Earth System Model properties and output from the University of Victoria
405 Earth System Climate Model (Keller et al., 2012, Mengis et al., 2021) including a) the model
406 bathymetry (depth levels), and b) the simulated present-day dissolved inorganic carbon
407 concentration (mol m^{-3}) averaged over the upper 50 m of the ocean. Panels c) and d) show
408 results from a coastal OAE study by Feng et al. (2017) where the change in upper ocean
409 alkalinity (upper 50 m) and the air-sea flux of CO_2 are shown relative to the RCP8.5 control
410 simulation. Shown is the Oliv100_Omega3.4 simulation from Feng et al. (2017), where 100 μm
411 olivine grains were added to ice-free coastal grid cells in proportion to RCP 8.5 CO_2 emissions
412 (i.e., 1 mol of alkalinity per mole of emitted CO_2) until a sea surface aragonite Ω threshold of 3.4
413 was reached.

414 Finally, another important challenge associated with global ocean models is the requirement to
415 represent the entire global ocean ecosystem with a single set of model parameters (e.g., Long et
416 al. 2021, Sauerland et al. 2020). In particular, the biological pump is an important control on the
417 distribution of biogeochemical tracers, including alkalinity and DIC. The magnitude of organic
418 carbon export, and the magnitude of biogenic calcium carbonate export, are important controls
419 on the distribution of alkalinity and DIC at the ocean surface and in the interior (e.g., Fry et al.,

420 2015). These quantities are a product of ecosystem function and, since the global ocean is
421 characterized by diverse biogeography (e.g., Barton et al., 2013), capturing global variations in
422 the biological pump presents a challenge.

423 2.1.4 Integration across scales

424 Choosing the appropriate modeling tool for a given OAE-related question requires clarity about
425 the scale of the problem to be addressed and the objectives of the model application.

426 Approaches for OAE vary significantly with respect to the spatial footprint of alkalinity
427 increase. Proposed methods for spreading alkalinity feedstocks at the surface ocean include the
428 addition of reactive minerals (e.g., CaO, Ca(OH)₂ or Mg(OH)₂) in ship-propeller washes (e.g.,
429 Köhler et al., 2013, Renforth et al., 2017, Caserini et al., 2021) or using other means (e.g., Gentile
430 et al., 2022) along tracks from commercial or dedicated OAE vessels or through coastal outfalls
431 (e.g., wastewater-treatment or power plants); the addition of less-reactive minerals to corrosive
432 or high-weathering environments (e.g., olivine spreading on beaches or mineral addition to
433 riverine discharge, e.g., Montserrat et al., 2017, Foteinis et al., 2023, Mu et al., 2023); and
434 electrochemically generated point-sources of alkalinity that are discharged as highly alkaline
435 seawater (e.g., House et al., 2009) from existing facilities (e.g., desalination and wastewater-
436 treatment plants), dedicated facilities (e.g., Wang et al., 2023), or from an array of smaller
437 infrastructure (e.g., grids of off-shore wind turbines). Models for OAE research should
438 represent these footprints of alkalinity increases appropriately for the questions being
439 addressed.

440 There are research questions that fall relatively neatly into one of the three scale ranges
441 described above in sections 2.1.1 to 2.1.3. For example, consideration of the nearfield effects of
442 different alkalinity feedstocks (e.g., dissolved versus particles) or analysis of the potential
443 impacts from secondary CaCO₃ precipitation due to elevated alkalinity from a point source
444 require models that resolve the scales of turbulent motion. Examination of the change in air-sea
445 CO₂ flux due to a broad and diffuse alkalinity increase is less demanding on model resolution
446 and regional scale models are appropriate for this question. Investigation of Earth system
447 feedbacks requires ESMs. However, there also are many aspects of OAE that require a bridging
448 of scales. For example, when considering different deployment methods like discharge from
449 vessels into the ocean surface boundary layer versus additions made through outfalls via
450 surface or subsurface plumes, modeling requirements vary. In both cases, the resulting
451 biogeochemical response may be affected by dynamics operating in the nearfield, where
452 conditions are sensitive to the deployment method and turbulence has to be considered, and the
453 far-field, where conditions do not depend on the details of how the alkalinity was added and
454 the air-sea flux of CO₂ is instead determined by ambient environmental processes. Another
455 example is the challenge that anomalies in air-sea CO₂ flux generated by OAE deployments will
456 manifest over large spatio-temporal scales because CO₂ equilibrates with the atmosphere via
457 gas exchange slowly. Some interplay among the modeling tools described in sections 2.1.1 and
458 2.1.2 is likely going to be required. One straightforward approach would be to parameterize
459 small-scale processes in the larger-scale models.

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472 **2.2 The range of biogeochemical realism & complexity**

473 Application of biogeochemical ocean models for the purposes of OAE research and verification
474 requires reevaluation, and likely further development, of several model assumptions and
475 features related to biogeochemical realism and complexity. For example, the internal sources
476 and sinks of alkalinity are typically not explicitly represented in ocean models; this may become
477 necessary in some circumstances but will be challenging (Section 2.2.1). OAE-related
478 perturbations of alkalinity, other carbonate system properties, and addition of macro- and
479 micronutrients contained in some alkalinity feedstocks may result in biological and ecosystem
480 responses that current biogeochemical models are not capable of representing but that would be
481 relevant for the assessment of environmental impacts of OAE and the verification its CDR
482 efficiency (Section 2.2.2). Furthermore, depending on the environmental setting, sediments can
483 be sources or sinks of alkalinity; these sediment-water fluxes need to be appropriately
484 considered, including the potential impacts of OAE on their magnitude, in order to obtain
485 complete and trustworthy carbon budgets (Section 2.2.3). Other boundary fluxes that require
486 accurate specification are alkalinity inputs from rivers and groundwater (Section 2.2.4) and the
487 air-sea flux of CO₂ across the air-sea interface (Section 2.2.5).

488
489 **2.2.1 Representing alkalinity in seawater**

490 Alkalinity is an emergent property that depends on the concentrations of numerous chemical
491 species with distinct internal source and sinks (Schulz et al. 2023; Wolf-Gladrow et al. 2007;
492 Middelburg et al. 2020). Skillful simulation of alkalinity in seawater may require explicit
493 representation of its multiple biotic and abiotic sources and sinks, some of which are difficult to
494 constrain. A major process by which alkalinity is consumed is the production of calcium
495 carbonate. In the water column, this is predominantly a biotic process, performed by calcifiers,
496 although “whiting” events, where calcium carbonate precipitates spontaneously from in
497 ambient seawater can be locally important (e.g., Long et al. 2017).

498
499 Models vary in the degree of mechanistic sophistication with which biogenic calcification is
500 represented. For example, some models explicitly resolve calcifiers, such as pelagic
501 coccolithophores (e.g., Krumhardt et al. 2017) and foraminifera (Grigoratou et al. 2022) and, in
502 some cases, also benthic corals, foraminifera, or calcifying higher trophic levels and thus can
503 mechanistically account for the associated alkalinity consumption. Alternatively, models can
504 parameterize biotic production of carbonate, and its subsequent sinking and dissolution, as a
505 fraction of organic matter production combined with an assumed remineralization profile (e.g.,
506 Schmittner et al. 2008; Long et al. 2021). Dissolution of carbonate minerals produces alkalinity,
507 at the sediment surface and in the water column as carbonate particles sink. This can be
508 represented with first-order abiotic dissolution kinetics with a dependence on the saturation
509 state of ambient water in the water column (e.g., Sulpis et al., 2021), in the sediments (e.g.,
510 Emerson & Archer, 1990) or in micro-environments in aggregates or organisms (Barrett et al.,
511 2014) with systematic differences for different crystal structures, aragonite and calcite (Morse et
512 al., 1980).

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520 Production of alkalinity occurs via uptake of nitrate or nitrite by photoautotrophs, while
521 remineralization consumes alkalinity when happening aerobically but generates alkalinity
522 when occurring anaerobically, e.g. via denitrification (Fennel et al. 2008). Biotic production and
523 consumption of alkalinity is stoichiometrically coupled to the release or uptake of nutrients and
524 carbon, where non-Redfield processes such as nitrogen fixation or denitrification need to be
525 specifically considered in the stoichiometric relationships (Paulmier et al., 2009).

526

527 Spontaneous precipitation of carbonate minerals in pelagic environments could occur when
528 seawater is highly oversaturated with respect to carbonate (Moras et al. 2022) but is, to the best
529 of our knowledge, not yet included in ocean models. When simulating OAE approaches that
530 may generate high oversaturation with respect to carbonate, spontaneous precipitation of
531 carbonates needs to be considered, especially when condensation nuclei are present.

532 Appropriate approaches will have to be developed, e.g., using near-field models to
533 mechanistically represent this process and a meta-model approach to develop
534 parameterizations that are suitable for far-field and larger-scale models.

535

536 Organic compounds produced within the ocean or originating from land can also act as proton
537 acceptors and contribute organic alkalinity (e.g., Koeve and Oschlies 2012, Ko et al. 2016,
538 Middelburg et al. 2020) and will impact the carbonate system, the partial pressure of CO₂ and
539 thus the air-sea CO₂ flux. Commonly, the contribution of organic alkalinity is deemed small
540 enough in oceanic environments to be negligible, but this assumption should be reconsidered in
541 the context of OAE, especially for coastal CDR deployments where the organic contribution to
542 alkalinity is thought to be larger. To the best of our knowledge, models do not account for
543 organic alkalinity. A better quantitative understanding of organic contributions to alkalinity is
544 likely needed to parameterize or mechanistically represent its contribution in models. Similarly,
545 it may be important in the context of mineral OAE deployments to account for local variations
546 in [Ca²⁺] and [Mg²⁺] to accurately estimate the pCO₂ anomalies generated by different OAE
547 feedstocks. While these constituents have very long residence times in the ocean, and are hence
548 commonly assumed to vary conservatively in proportion to salinity, variations in their relative
549 abundance has an impact on the thermodynamic equilibrium coefficients used to solve seawater
550 carbonate chemistry (Hain et al., 2015).

551

552 **2.2.2 Representing biological and ecological processes**

553 A key question related to OAE is whether changes in carbonate chemistry induce differential
554 responses in organisms. In the pelagic zone, OAE might shift the phytoplankton community
555 composition, for example, due to distinct physiological sensitivities of different groups (e.g.,
556 Ferderer et al. 2022). Further, if OAE is accomplished via rock dissolution, carbonate versus
557 silicate rock may impact the relative balance between phytoplankton functional groups (PFTs)
558 such as calcifiers and diatoms, and changes in Mg and Ca ratios may also influence calcification
559 (Bach et al., 2019). Additionally, ancillary constituents specific to particular feedstocks may have
560 biological activity. Silicate rocks include bioreactive metals such as Fe, a micronutrient with the
561 capacity to stimulate phytoplankton growth, and others that are can be toxic when occurring in

562 high concentrations, such as Ni and Cu, and may adversely impact phytoplankton and reduce
563 primary productivity (Bach et al., 2019). The bioreactivity of these metals may be difficult to
564 simulate in models as their dissolved concentrations can be partially mediated by complexation
565 with organic ligands (Guo et al., 2022). Physical impacts of OAE feedstocks may also have
566 important biological impacts through changes in the propagation of light in the surface ocean,
567 and direct exposure to mineral particles may have additional impacts, e.g., on zooplankton
568 through particle ingestion (Harvey, 2008; Fakhraee et al., 2023). Effects of OAE on plankton
569 have the potential to propagate to higher trophic levels through marine food webs as the
570 magnitude and quality of net primary productivity shifts and trophic energy transfer is altered
571 accordingly.

572
573 Simulating this full collection of processes in models is challenging. Dominant modeling
574 paradigms for simulating planktonic ecosystems include PFT- and trait-based models (e.g.,
575 Negrete-Garcia et al., 2022). In these systems, physiological sensitivities are parameterized
576 according to transfer functions that modulate rate processes—growth, for instance—on the basis
577 of ambient environmental conditions. Nutrient limitation of growth is often represented using
578 Michaelis–Menten kinetics wherein growth rates decline as nutrients concentrations become
579 limiting. State-of-the-art ESMs represent PFTs with multiple nutrient co-limitation, which is
580 essential to effectively simulate plankton biogeography of the global ocean. Diatoms, for
581 example, are capable of high growth rates, enabling them to outcompete other phytoplankton
582 under high-nutrient conditions, but their range is restricted to high latitudes and upwelling
583 regions where there is sufficient silicate. If OAE were to modulate the concentration of
584 constituents represented by multiple nutrient co-limitation models, it is possible such models
585 could simulate the phytoplankton community response—though it’s important to consider
586 whether the models provide representations that are sufficiently robust for the magnitude of
587 OAE-related perturbations. In some cases, models are missing key processes that would be
588 required to mechanistically simulate certain effects. We are aware of no models that represent
589 Ni toxicity, for instance. Including these effects, as well as a capacity to simulate secondary
590 interactions, such as ligand complexation of metals in OAE feedstocks, will require significant
591 investment in empirical experimentation to understand essential rate processes and
592 physiological responses.

593
594 Shortcomings in the capacity of models to represent physiological responses to OAE is an
595 important consideration for the ability of models to faithfully represent ecological impacts.
596 Notably, electrochemical OAE techniques present a simpler set of processes to consider than
597 using crushed-rock feedstocks, where ancillary constituents and physical dynamics come into
598 play. For electrochemical OAE, the most likely biological feedback to consider relates to the
599 impacts of changing carbonate chemistry on biogenic rates of calcification or phytoplankton
600 growth rates (Paul and Bach 2020). It is also possible that carbon limitation of phytoplankton
601 growth (Paul and Bach 2020; Riebesell et al. 1993) may also be important. Empirical research
602 exploring physiological sensitivities should be used to develop prioritizations of key model
603 processes comprising early targets for implementation. Model documentations should use
604 consistent stoichiometric relations to link alkalinity changes to those of nutrients and carbon

605 (Paulmier et al. 2009) and state the assumptions made about carbonate formation and
606 dissolution.

607

608 2.2.3 Representing sediment-water exchanges

609 The exchange of solutes between the sediments and overlying water influences ocean
610 chemistry, including the properties of the carbonate system (Burdige 2007). Depending on
611 location and time scale, OAE may affect these exchanges and should be appropriately
612 considered in models. Sediments influence the marine carbonate system primarily through the
613 remineralization of organic matter, which returns DIC to overlying water (and alkalinity if this
614 remineralization occurs anaerobically), and the dissolution of biogenic silicate or carbonate
615 minerals. CaCO_3 is of particular importance as its dissolution releases alkalinity, while its burial
616 is an alkalinity sink, and the balance between the two is a key control on the ocean's alkalinity
617 balance over timescales approaching 10^4 years (Middelburg et al. 2020). Furthermore,
618 remineralization and other microbial metabolisms, such as "cable bacteria," can significantly
619 lower pore water pH by several pH units below seawater values (Meysman and Montserrat
620 2017). This can drive dissolution of CaCO_3 and generate alkalinity in the sediments, even in
621 shallow waters when the overlying water is supersaturated (Rau et al. 2012).

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622
623 Representing these processes in coastal and shelf sediments (< 200 m) is challenging. Shallow
624 water depths and high productivity result in a significant delivery of organic matter to the
625 sediments that is much larger than in the deep ocean. As a result, the relative importance of
626 sediments in organic matter remineralization is larger and production of alkalinity by anaerobic
627 metabolisms is more important in these shallow sediments than in the deep ocean (Seitzinger et
628 al. 2006, Jahnke 2010, Huettel et al. 2014, Chua et al. 2022). In addition, these environments are
629 dynamic with organic supply and bottom water conditions varying on tidal, seasonal, and
630 interannual timescales. Accounting for the exchange between sediments and overlying water
631 and its variability on tidal, seasonal, and interannual timescales will likely be necessary in
632 regional and global biogeochemical models that aim to simulate alkalinity cycling in coastal and
633 shelf seas, even for relatively short simulation durations of months to years.

634
635 The choice of approach to modeling sediments may depend on the sediment type. For example,
636 the mechanisms transporting solutes across the sediment-water interface can be divided into
637 two categories depending on the sediment's grain size. In coarse sediments, i.e. permeable
638 sands, pressure gradients drive flow through the seabed replenishing sediment oxygen content
639 (Huettel et al. 2014). Organic carbon stores are low and remineralization was long thought to be
640 primarily aerobic. However, evidence has emerged relatively recently that anaerobic
641 remineralization in sandy sediments is more important than originally thought (Chua et al. 2022
642 and references therein). Idealized models that represent the three-dimensional sediment
643 structure illustrate the importance of turbulence and oscillatory flows in permeable sediments
644 (see Box 2 in Chua et al. 2022). These models are highly localized and computationally
645 demanding, prohibiting their coupling with ocean biogeochemical models. Thus, permeable

649 sediments are currently not well represented in regional or global ocean biogeochemical
650 models.

651
652 In cohesive, fine-grained sediments with low permeability, i.e. muds, transport is limited by
653 diffusion or faunal mediated mixing and exchange processes, i.e. bioirrigation or bioturbation
654 (Meysman, et al. 2006, Aller 2001). In these environments, detailed multicomponent reactive-
655 transport models of sediment biogeochemistry – so called diagenetic models – can reproduce
656 carbon remineralization rates partitioned between aerobic and anaerobic pathways,
657 precipitation/dissolution reactions between sediment grains and porewaters, and the transport
658 of solutes across the sediment-water interface (Boudreau 1997, Middelburg et al., 2020). These
659 mechanistic models will be useful for detailed investigations into how perturbations of the
660 carbonate system in seawater overlying the sediments affect their biogeochemistry and for
661 addressing questions about the potential influence of particulate alkalinity feedstocks settling to
662 the seafloor (Montserrat et al. 2017, Meysman and Montserrat 2017). However, typically these
663 models are one-dimensional and applied to a few representative locations. Coupling fully
664 explicit diagenetic models to three-dimensional ocean biogeochemical models, while
665 conceptually straightforward, is computationally prohibitive. Instead, depth-integrated
666 sediment processes have been implemented as bottom boundary conditions (e.g., Moriarty et al.
667 2017, 2018, Laurent et al. 2016). For example, Laurent et al. (2016) used a diagenetic model in a
668 “meta-modeling” approach to estimate bottom boundary nutrient fluxes for a regional scale
669 biogeochemical model. By parameterizing the diagenetic model with detailed geochemical data
670 (porewater profiles and nutrient fluxes) from a few individual locations, then forcing it over a
671 range of expected bottom water conditions, they developed empirical functions relating
672 sediment fluxes to bottom water conditions that could be used to parameterize bottom
673 boundary conditions in the water column model. A similar approach could be used in OAE
674 models to parameterize how sediment biogeochemistry may alter alkalinity fluxes, for example,
675 how redox sensitive processes, such as coupled nitrification-denitrification or sulfate reduction
676 coupled to pyrite burial, both of which may produce alkalinity (Soetaert et al. 2007), may
677 respond to changes in bottom water oxygen or organic matter loading.

678
679 When considering the long-term storage of CO₂ in global-scale ESMs, the interactions between
680 sediments and the deep ocean (> 1000 m bottom depth) may need to be considered. In this
681 environment most organic matter remineralization occurs in the water column, and the small
682 amount of organic matter reaching the seafloor is remineralized aerobically with little to no
683 release of alkalinity. In this case, sediment remineralization can likely be either ignored or
684 implemented as a reflective boundary condition where the simulated POC flux to the seafloor is
685 immediately returned as DIC and remineralized nutrients. However, the dissolution or
686 preservation of CaCO₃ in deep sediments is critical to controlling deep water alkalinity and may
687 be important in model simulations that aim to quantify OAE effects on the timescales associated
688 with the large-scale global overturning circulation. CaCO₃ solubility increases with pressure
689 and [decreasing](#) pH and [CaCO₃](#) eventually becomes undersaturated at depth. The depth at
690 which sinking CaCO₃ balances its dissolution is referred to as the carbonate compensation
691 depth (CCD). An increase in bottom water CO₃²⁻ or CaCO₃ deposition, will deepen the CCD,

692 burying CaCO₃, trapping alkalinity, and lowering the alkalinity budget of the ocean.
693 Conversely if CaCO₃ rain rate or CO₃²⁻ concentration decreases, the CCD will shoal and
694 previously buried CaCO₃ will dissolve releasing alkalinity to the deep ocean. CCD
695 compensation therefore opposes any forcing of the deep ocean carbonate system and therefore
696 dampens the rise of CO₂ in the atmosphere but will also counteract any potential OAE solution
697 (see Renforth and Henderson 2017 for a detailed explanation). Although most CaCO₃
698 dissolution occurs in the sediments, there is no consensus as to the level of detail this needs to
699 be represented in models. Some global models employed to investigate large-scale OAE include
700 calcium carbonate dynamics at the sediment surface (Ilyina et al. 2013) others disregard this
701 process (Keller et al. 2014).

702
703 Often global models will parameterize CaCO₃ burial as a function of saturation state, such an
704 approach is effective for resolving CCD dynamics over geological timescales (~10,000 y), but not
705 over the century to millennial timescales of CCD readjustment. Models that fully couple
706 sediment diagenesis can resolve these dynamics (Gehlen et al. 2008), but the computational
707 demand can make them ineffective. One solution is the approach of Boudreau et al. (2010) and
708 (2018). By suggesting that CaCO₃ dissolution dynamics are controlled by transport of
709 dissolution products across the benthic boundary layer, they were able to derive equations
710 predicting CCD depth and CaCO₃ dissolution based on bottom water CO₃²⁻ and CaCO₃ rain rate
711 and avoiding a detailed representation of the sediments. These equations, combined with model
712 bathymetry, can parameterize sediment CO₃²⁻ flux as a boundary condition and suitably account
713 for transient sediment CaCO₃ dissolution in large scale ESMs while avoiding the computational
714 demands of a fully coupled ocean circulation-diagenesis model.

715 716 *2.2.4 Representing river and groundwater fluxes*

717 Regional and global ocean biogeochemical models typically account for river inputs, including
718 their contributions to alkalinity and **DIC**. In most models this is done by specifying alkalinity
719 and **DIC** concentrations in imposed riverine freshwater fluxes, although accurate prescription of
720 these concentrations can be challenging. Typically, a combination of direct river measurements,
721 where available, output from watershed models (e.g., Seitzinger et al. 2010), or extrapolations of
722 coastal ocean measurements to a freshwater endmember (e.g., Rutherford et al. 2021) are used.
723 Solute inputs from groundwater are typically ignored but could be important locally. In high-
724 resolution coastal domains near urban areas, sewage input may be an additional important
725 source of carbon, nutrients, and alkalinity.

726
727 It is important to note that land-based CDR applications may have an important effect on ocean
728 alkalinity dynamics through riverine and groundwater delivery of solutes. Terrestrial OAE
729 equivalents broadly referred to as Enhanced Rock Weathering (ERW) rely on the application of
730 lime or pulverized silicate or carbonate rocks on land and in rivers. These strategies aim to
731 generate CO₂ uptake locally but yield a leaching flux of bicarbonate into freshwater systems and
732 subsequent transport into the coastal ocean. Field trials and some commercial applications are
733 currently underway, most of them with the implicit or explicit assumption that the enhanced

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736 delivery of alkalinity will generate a carbon removal in the ocean (Köhler et al., 2010; Taylor et
737 al., 2016; Bach et al., 2019). There is a need for coordinated efforts to improve quantification of
738 background riverine fluxes and establish initiatives to effectively track the solute additions from
739 ERW.

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741 2.2.5 Representing air-sea gas exchange

742 The calculation of air-sea gas exchange is necessary for the quantification of net carbon uptake
743 from OAE in models. Biogeochemical models typically represent this exchange using a bulk
744 relationship that depends on the product of the gas transfer velocity and the effective air-sea
745 concentration difference (Fairall et al. 2000). However, the gas transfer velocity remains highly
746 uncertain and is sensitive to a collection of processes that vary across scales, including sea state,
747 boundary layer turbulence, bubble dynamics, and concentrations of surfactants. The most
748 widely used parameterizations of the gas transfer velocity use empirical fits to observations to
749 construct a functional relation dependent on wind speed only, under the premise that
750 turbulence and bubbles (via the breaking of surface gravity waves) are predominantly
751 determined by wind stress (Wanninkhof 2014). This neglects processes that could be regionally
752 important such as convection, modification by biological surfactants, rain and wave-current
753 interactions, while vastly simplifying the effects of wave breaking and bubbles. Although
754 different dependencies on wind speed have been proposed (quadratic, cubic, hybrid),
755 parameterizing the gas transfer coefficient as a quadratic function of the 10-meter wind speed is
756 the most common (Wanninkhof 1992; Wanninkhof 2014). This relationship is supported by
757 direct measurements of air-sea flux at intermediate wind speeds (3-15 m/s), but at low wind
758 speeds (< 3 m/s), non-wind effects can have an important impact on gas transfer. At high wind
759 speeds (> 15 m/s), breaking waves and bubble injection enhance gas exchange for lower
760 solubility gasses such as CO₂ (Bell et al. 2017). Therefore, quadratic fits tend to underestimate
761 the gas exchange at low and high wind speeds (Bell et al. 2017).

762 More complex air-sea exchange parameterizations account for processes such as bubbles, near
763 surface gradients and buoyancy driven convection (e.g., Liang et al. 2013, Fairall et al. 2000), but
764 they depend upon a wider range of input variables. Other considerations in estimating flux
765 arise from the nonlinear dependence on these variables, e.g., wind speed, which can lead to
766 underestimates when made using daily averages rather than hourly measurements (Bates and
767 Merlivat 2001).

769 Notably, the gas transfer velocity (k_w) determines the kinetics of gas exchange, given a
770 perturbation in surface ocean $p\text{CO}_2$ away from equilibrium. The timescale for CO₂ equilibration
771 over the surface mixed layer can be fully quantified using the following expression,

$$772 \tau_{gas-ex} = \left(\frac{\partial \text{CO}_2}{\partial \text{DIC}} \right)^{-1} \left(\frac{h}{k_w} \right)$$

774 where h is the depth of the surface mixed layer and the partial derivative $\partial \text{CO}_2 / \partial \text{DIC}$ captures
775 the thermodynamic state of the carbon system chemistry in seawater, specifically with respect
776 to the amount that dissolved CO₂ changes per unit change in DIC (Sarmiento and Gruber 2006).
777 This property is related to the buffer capacity and varies in roughly linear proportion to the

779 carbonate ion concentration. The magnitude of $\left(\frac{\partial \text{CO}_2}{\partial \text{DIC}}\right)^{-1}$ is typically about 20, which explains
780 why the equilibration timescale for CO₂ is so long. The contribution of uncertainty in the gas
781 exchange velocity to overall uncertainty in carbon uptake from OAE deployments will depend
782 in part on the circulation regime involved. For example, in situations where alkalinity-enhanced
783 water parcels are retained at the surface for timescales that are significantly longer than $\tau_{\text{gas-ex}}$,
784 full equilibration will occur and the impact of uncertainty in the gas exchange velocity will have
785 limited influence on the overall uncertainty.

786
787 Even though OAE-induced additional air-sea CO₂ fluxes will, even in hypothetical massive
788 deployments, amount to at most a few Gt CO₂/yr, which is typically not more than a percent of
789 the atmospheric CO₂ inventory, this subtle difference in the treatment of the atmospheric
790 boundary condition can be significant. Using prescribed atmospheric $p\text{CO}_2$ that is unresponsive
791 to marine CDR-induced air-sea CO₂ fluxes has been shown to overestimate oceanic CO₂ uptake
792 by 2%, 25%, 100% and more than 500% on annual, decadal, centennial, and millennial
793 timescales, respectively (Oschlies, 2009). Simulations with prescribed atmospheric $p\text{CO}_2$ need to
794 take such systematic biases into account.

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796 2.3 Model development needs for OAE research

797 While there is already substantial capacity for simulating ocean biogeochemical dynamics at
798 global to regional scales, the discussion above implicates several areas where additional efforts
799 are required to fully establish a modeling capability suitable for supporting OAE. These fall into
800 four primary areas: (1) supporting multi-scale simulations with sufficiently high-fidelity flow
801 fields; (2) faithfully simulating the near-field dynamics associated with alkalinity addition; (3)
802 capturing feedbacks to OAE owing to biological and geochemical responses; and (4) identifying
803 whether there are reduced-complexity modeling approaches that might provide sufficiently
804 robust estimates of the net effects of OAE.

805 As elucidated above, a primary consideration related to capturing OAE impacts is the fidelity of
806 the simulated flow. Notably, OAE presents a somewhat novel use case requiring an effective
807 multi-scale modeling capability. A conceptually straightforward path to improving the
808 representation of ocean circulation and mixing is to increase the resolution of the model grid.
809 However, the computational demand of high-resolution simulations can only be met over more
810 limited-area domains. Since the spatiotemporal footprint of OAE-related perturbations is likely
811 to be large, there will be a need to represent large regions. An argument might be made,
812 however, that the circulation in proximity of an OAE site is most important to capture with
813 high-fidelity. This can be achieved with two-way nested regional models as described in see
814 Section 2.1.2 but will require further development to couple in the nearfield models described in
815 Section 2.1.1. Native grid-refinement, e.g. via unstructured grids, is another approach that may
816 be pursued to effectively support OAE research.

817 The second area of model development relates to the requirement of faithfully representing the
818 dynamics associated with alkalinity addition. Regional to global scales are the most relevant for
819 simulating the air-to-sea exchange of CO₂ ensuing from OAE. It is important, however, to

822 ensure that local processes affecting the mass fluxes and initial dispersal of alkalinity are
823 handled appropriately. As illustrated above, DNS or LES simulations (section 2.1.1) can be
824 leveraged to develop parameterizations for larger-scale models, including for crushed-rock
825 feedstocks where particle dynamics may be important or techniques involving alkalinity
826 enhanced streams entering the ocean from outfall pipes. In addition to process fidelity, there are
827 also numerical constraints to consider. For example, advection schemes used in most ocean
828 general circulation models struggle to represent sharp gradients; large mass fluxes of alkalinity
829 into single model grid-points are likely to cause advection errors that may contaminate aspects
830 of the model solutions making interpretation difficult. More specifically, conservative advection
831 schemes can be characterized in terms of their accuracy, monotonicity (i.e., ability to preserve
832 sign), and linearity (i.e., ability to preserve additivity) and there are always tradeoffs to make
833 between these properties. Research may be required to determine which schemes are best
834 suited to the particular challenges associated with representing the advection of OAE signals.

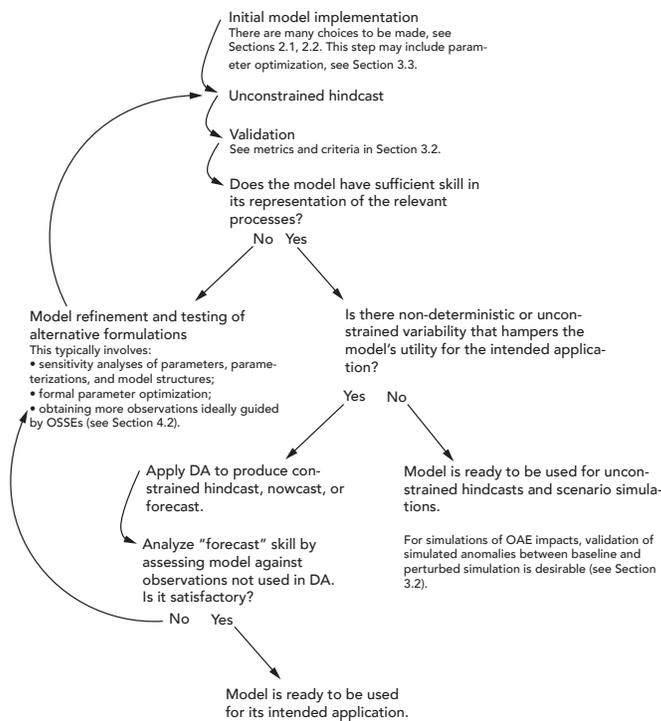
835 The third area of model development relates to our capacity to fully capture the range of
836 biogeochemical feedback associated with OAE. The class of processes to consider here is
837 potentially large and many have been touched on in section 2.2.1 to 2.2.3. Precipitation
838 dynamics, specific elemental components of alkalinity, biogenic responses mediated by
839 physiological or ecological sensitivities, impacts and processes controlling the cycling of
840 ancillary constituents, and accurate sediment-water exchange are all areas that merit
841 consideration. Further efforts are required to understand and prioritize these areas of potential
842 development and, notably, their relative importance is likely to be regionally dependent.

843 Finally, it is important that models be tailored to address specific questions of relevance. In this
844 context, it may be important to consider how much model complexity is required to capture the
845 effects of perturbations, seeking parsimonious representations that are well-supported by
846 empirical constraints and invoking wherever possible a separation of concerns to isolate the
847 factors contributing to uncertainty. For example, there are several near-field considerations that
848 might be addressed using a combination of local observations and ultra-high-resolution
849 modeling tools to generate estimates of alkalinity mass fluxes that are subsequently imposed as
850 forcing in regional- to global-scale models. Another key question is how important it is to
851 comprehensively simulate the mean state to faithfully capture the response to OAE
852 perturbations [for the purpose of MRV](#). For example, if it can be documented that biological
853 feedbacks to OAE are of negligible concern, the core target for simulating OAE effects [for MRV](#)
854 may be to capture the cumulative integral of air-sea CO₂ exchange associated with the induced
855 surface ocean pCO₂ anomaly. The mean state of the seawater carbon system is relevant here as
856 the background DIC and alkalinity fields determine the pCO₂ response per unit addition of
857 alkalinity, but fully prognostic calculations of nutrient cycling may not be necessary.

858 **3 Model validation and integration with observations**

859 Whether a model is useful for OAE research depends on how accurately it represents the
860 physical, chemical, and biological processes that are relevant to the specific research question to
861 be addressed. Model validation, the evaluation of a model's performance, and estimation of

862 uncertainties in model output should thus be integral parts of model implementation and
 863 application. It is important to note that any model, even after best efforts have been made to
 864 improve formulations and conduct the most thorough validation, will deviate from reality. Any
 865 model is, by definition, a simplification of the real world and thus its output will be subject to
 866 uncertainties. Deviations of the model state from the real world can be reduced by applying
 867 statistical techniques, collectively referred to as Data Assimilation (DA) methods, that combine
 868 models with observations and yield the best possible estimates. The steps typically involved in
 869 model implementation and validation, and possible integration with observations through data
 870 assimilation are shown in Figure 3. In this section, we summarize the most important
 871 observation needs for model validation (Section 3.1), briefly describe typical metrics for model
 872 validation and articulate a reasonable minimum criterion (Section 3.2), give a high-level
 873 explanation of approaches for the formal statistical combination of models with observations
 874 through parameter optimization and state estimation (Section 3.3), and describe approaches for
 875 the specification of uncertainty in model outputs (Section 3.4).
 876



877
 878
 879 **Figure 3:** Typical steps in model implementation and validation.

880 **3.1 Observation types for validation**

881 Two fundamental requirements for models to be useful in the context of OAE research are high-
882 fidelity representations of physical transport due to advection and mixing, and of
883 biogeochemical effects of OAE, most importantly changes in the inorganic carbon properties.

884 Observations for validation of the simulated physical transport of alkalized waters include
885 temperature and salinity distributions, direct measurements of currents, surface drifter
886 trajectories, sea surface height observations from satellite altimetry, and estimates of
887 geostrophic flow derived from the latter. Additional metrics relevant for assessing the fidelity of
888 the large-scale overturning circulation in global models include combinations of biogeochemical
889 concentration and transient tracers. For example, oxygen can be useful for identifying large-
890 scale transport pathways, even though it convolutes dynamical and biological information.
891 Particularly valuable for assessing large-scale ocean transport on the timescales relevant for
892 OAE are abiotic transient tracers such as such as chlorofluorocarbons (CFCs), sulfur
893 hexafluoride (SF₆), and possibly the isotopes ³⁹Ar and ¹⁴C. Observational approaches for
894 validation at regional scales include explicit tracer studies for documenting dispersion
895 properties using Rhodamine dye or SF₆.

896 In addition to the dynamics of the flow, model validation for OAE research requires the
897 assessment of the fidelity of simulated carbonate chemistry variables (e.g., alkalinity, total
898 dissolved inorganic carbon or DIC, pH, pCO₂) and salinity and temperature, which are used to
899 calculate the 13 thermodynamic equilibrium constants and conservative chemical species
900 needed to constrain seawater acid-base chemistry in oxygenated seawater. Depending on the
901 OAE approach and the model application, assessment may also require observed macronutrient
902 (e.g., nitrate, silicate, or phosphate), micronutrient (e.g., Fe), and contaminant (e.g., Ni, and Cr)
903 measurements; bulk seawater properties related to biogeochemical cycling (e.g., dissolved
904 organic carbon content [DOC], particulate inorganic carbon [PIC], chlorophyll fluorescence);
905 and biogeochemical rates and fluxes (e.g., net community calcification).

906 It is not always feasible to obtain the ideal carbonate system observations for model validation.
907 Temperature and salinity can be measured reliably across all ocean depths and, with greater
908 uncertainty and only at the ocean surface, remotely from satellites. The technical capacity for
909 seawater pH measurements is evolving rapidly and sensors and systems now exist for pH
910 measurements across nearly all depths, though the depth-capable systems require regular
911 recalibration (e.g., Maurer et al., 2021). Similarly, there are numerous ways to observe surface
912 ocean pCO₂ using a variety of crewed, autonomous, and fixed-location platforms (e.g., ship-
913 based, Saildrone, and moored systems). However, interior-ocean pCO₂ observations remain
914 challenging to obtain due to the need for calibration gasses and a gas-water interface. Alkalinity
915 titrations are predominantly performed on discrete bottle samples collected by hand, though
916 autonomous titration systems are under development that enable *in situ* surface time series
917 measurements (Shangguan et al., 2022). Microfluidic *in situ* alkalinity titrators are also under
918 development that consume less reagent per sample but currently show higher uncertainties
919 than discrete samples (Sonnichsen et al. 2023). Solid state titrators that generate acid titrant *in*

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923 *situ* show promise for surface and subsurface alkalinity titrations, but these sensors are still
924 undergoing development and validation (Briggs et al., 2017). DIC observations combine the
925 limitations of current measurement systems for both the $p\text{CO}_2$ and alkalinity, and there are only
926 a handful of automated DIC titration systems rated for surface ocean measurements (e.g.,
927 Fassbender et al. 2015; Wang et al. 2015; Ringham 2022). Theoretically, measurement of two of
928 the carbonate system parameters in combination with temperature and salinity and some
929 additional assumptions allows calculation of the other carbonate system parameters in
930 seawater. Unfortunately, the pair of $p\text{CO}_2$ and pH, which are the most accessible to autonomous
931 measurement among the carbonate system parameters, provide nearly identical information
932 about the system. Thus, the results of the calculations that use this pair have higher
933 uncertainties than other combinations (Dickson and Riley 1979; Millero 2007; Cullison Gray et
934 al. 2011; McLaughlin et al. 2015; Raimondi et al. 2019) and are therefore not ideal as a pair for
935 model validation.

936 3.2 Validation metrics and approach

937 Validation relies on comparing the model output to observations, often in an iterative loop
938 where the evaluation of a hindcast simulation is followed by model refinements followed in
939 turn by a new hindcast and re-evaluation (Figure 3, Rothstein et al. 2006). Several evaluation
940 metrics are commonly used (see Box 3 in Fennel et al. 2022). The three most common are the
941 root-mean-square error (RMSE), the bias, and the correlation coefficient. All three are relative
942 measures without any objective criterion that indicates which range of values is acceptable or
943 unacceptable. In contrast, the Z-scores, which consider variability within the observational data
944 set, and the so-called model efficiency or model skill, which quantifies whether the model
945 outperforms an observational climatology are two metrics with built-in criteria as to whether a
946 model's performance is acceptable or not (Fennel et al. 2022). Since no single metric provides a
947 complete picture of a model's skill, multiple complementary metrics should always be used in
948 combination (Stow et al. 2009). Furthermore, different points in space and time, and a breadth of
949 variable types should be part of any comprehensive validation because a model may provide
950 accurate estimates for some variables, locations, or times but perform poorly for others (Doney
951 et al. 2009).

952 For OAE research, validation can be considered as a two-step challenge. First, it is necessary to
953 validate unperturbed model baselines to gain confidence that the natural variability is
954 represented appropriately and to quantify model uncertainties. One should compare model-
955 simulated spatial fields and time-series at strategic locations with appropriate observations to
956 assess the model's skill at representing mean distributions as well as the variability for
957 carbonate chemistry measurements and other relevant properties using several of the
958 complementary quantitative metrics listed above. A model could be considered as sufficiently
959 validated when mean distributions, their seasonal variability, and the timing and magnitude of
960 events (e.g., blooms, physical disturbances) are accurately represented. As described in Section
961 3.1, insufficient availability of observational constraints on carbonate system parameters
962 presents a major challenge in this regard. In models applied for OAE research, it is particularly
963 important to assess whether they realistically capture the distributions and variability of

964 seawater properties that govern sensitivity of the seawater carbonate system; recent work by
965 Hinrichs et al. (2023) shows that the current representation of alkalinity in state-of-the-art
966 models requires improvements.

967 The second, even more difficult step is to test whether a model accurately represents alkalinity
968 additions. OAE-related modeling studies thus far have relied on models that are validated only
969 for baseline conditions. These are useful as sensitivity studies. However, validation of a model's
970 ability to accurately represent the perturbations of an alkalinity addition is ultimately needed to
971 address OAE science questions around environmental impacts and MRV. It is likely that the
972 metrics described above for baseline validation are not suitable for this task. Validation should
973 focus on quantifying whether the model accurately captures the anomalies created by OAE.
974 This requires consideration of the spatial footprint and temporal evolution of perturbations and
975 ideally a close integration of experimental, observational, and modeling efforts. For example, a
976 model that is deemed skillful after baseline validation can be used to estimate the appropriate
977 dosage of alkalinity additions, thus ensuring a measurable signal, and guide the observational
978 strategy; subsequent validation may indicate model shortcomings that were not obvious in the
979 baseline validation (e.g., diverging dissipation rates between model and field observations) and
980 prompt model refinement in an iterative loop of model validation, improvement, and renewed
981 experimental assessment (Figure 3).

982 It is important to note that even with repeated steps of validation and model improvement,
983 there is going to be a limit to the degree of realism that can be achieved with any model. Any
984 model simulation will be prone to errors and uncertainties. Sources of error include inaccuracies
985 in model inputs, numerical approximation schemes, insufficient process understanding, and
986 inaccurate model parameters and parameterizations.

987 3.3 Data Assimilation

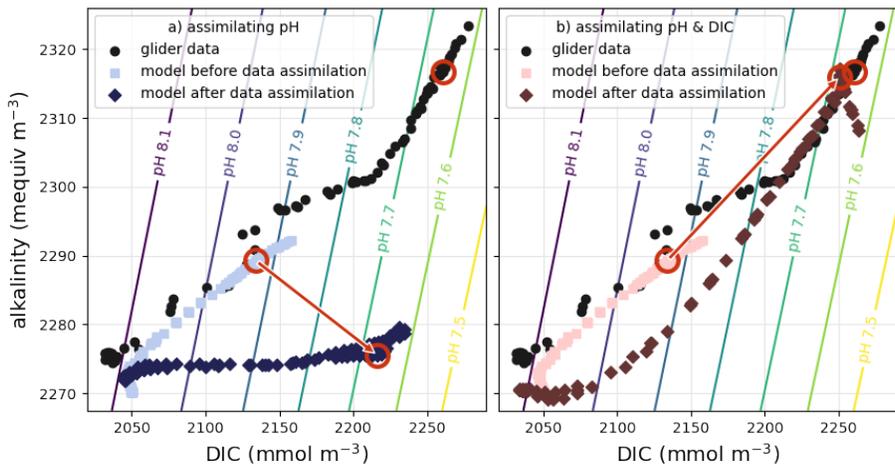
988 Data assimilation (DA) is the process of improving the dynamical behavior of models by
989 statistically combining them with observations. There are a variety of DA techniques that rely
990 on different mathematical and statistical approaches (Carrassi et al. 2018). Originally developed
991 for numerical weather prediction, DA has been successfully applied to ocean models, including
992 biogeochemical models (Mattern et al. 2017, Cossarini et al. 2019, Ciavatta et al. 2018, Verdy and
993 Mazloff 2017, Teruzzi et al. 2018, Fennel et al. 2019) but success critically depends on the
994 information content of the available observations (Yu et al. 2018; Wang et al. 2020). While DA
995 has been shown to yield large improvements in important parameters governing
996 biogeochemical processes (Mattern et al. 2012, Schartau et al. 2017, Wang et al. 2020) and in
997 model estimates of the physical and biogeochemical model state (Hu et al. 2012, Mattern et al.
998 2017, Ciavatta et al. 2018), it is only starting to be applied to carbonate system properties (Verdy
999 and Mazloff 2017, Carroll et al. 2020, Turner et al. 2023, Figure 4).

1000 Application of DA for ocean models is typically applied for one of two purposes: (1) to
1001 systematically optimize model parameters, e.g., phytoplankton growth and nutrient uptake or
1002 rates of background dispersion, and (2) to estimate [the](#) ocean state, e.g., distributions of
1003 temperature, phytoplankton biomass, alkalinity (see Fennel et al. 2022 for more details on the

1004 practical approaches and examples). The first purpose addresses systematic errors and biases in
 1005 models and is useful when systematically modifying and testing different model formulations
 1006 while the second assumes an unbiased model and addresses unresolved stochasticity, e.g.,
 1007 correcting the locations of mesoscale eddies and current meanders. State estimation offers the
 1008 potential to constrain variability such that OAE-induced perturbations of carbonate system
 1009 parameters can be documented even if they are smaller than the natural variability in the study
 1010 region. Joint estimation of physical and biogeochemical properties is common and can yield
 1011 significant improvements for both types of properties (Yu et al. 2018). Hybrid approaches
 1012 combining parameter and state estimation have also been proposed (Kitagawa 1998, Mattern et
 1013 al. 2012, 2014) but are less widely used.

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1015
 1016 **Figure 4:** Example of a DA application for state estimation of carbonate system properties
 1017 within a 3-dimensional model of the California Current System. The symbols show glider data
 1018 and model estimates at the measurement times and locations; one specific data point and its
 1019 associated model estimates are highlighted by red circles. Each data point consists of measured
 1020 pH alongside estimated alkalinity and DIC values (see Takeshita et al. (2021) for data source
 1021 and details). In the model, pH is a diagnostic variable and primarily dependent on the model's
 1022 alkalinity and DIC estimates. (a) When only pH data is assimilated, the model estimates are
 1023 moved closer to the observed pH values by increments in alkalinity-DIC space that degrade the
 1024 model's alkalinity estimates. (b) The model state estimates improve considerably by
 1025 assimilating data for DIC (or alkalinity; not shown) together with the pH observations.

1026 Successful application of DA critically requires sufficient observations either of the properties
 1027 that the model parameters to be estimated depend on or of the state variables that are being
 1028 estimated. The most commonly used observation type in biogeochemical DA applications is
 1029 satellite-based ocean color observations (Mattern et al. 2017, Ciavatta et al. 2018, Teruzzi et al.

1031 2018) which are available at a relatively high temporal resolution and covering large areas of the
1032 surface ocean. While these observations are useful for informing model estimates of properties
1033 directly linked to processes involving phytoplankton, they provide little information on the
1034 carbonate system. Dynamical models are able to quantitatively constrain processes that cannot
1035 be measured directly, by inferring them from observable properties, but only if the observations
1036 contain enough relevant information about the processes of interest. Hence, one of the biggest
1037 challenges facing the application of DA to models of the marine carbonate system, is the
1038 sparsity of observations of the marine carbonate system. Observations of pH, pCO₂, alkalinity,
1039 and DIC used to be limited to moorings and research cruises but have more recently been
1040 extended by automated observing systems, such as gliders, BGC-Argo floats and uncrewed
1041 surface vehicles (Bushinski et al. 2019). Although these measurements are becoming more
1042 common (Chai et al. 2020), they are still sparse compared to what is typically required for DA
1043 applications. In this context, an additional challenge is the problem of underdetermination, i.e.
1044 if multiple processes or properties of interest can cause a similar change in an observable
1045 property, then observing this property alone may not hold enough information to constrain
1046 these processes or properties and more observations are needed (see Fig. 4 and code examples
1047 in Fennel et al. 2022). As new platforms are added to the observing system, DA techniques can
1048 help guide their optimal deployment and tailor observational programs to the specific needs of
1049 OAE applications (see Section 4.3 below). Furthermore, statistical and machine-learning
1050 approaches are being developed (e.g., Lohrenz et al. 2018, Bittig et al. 2018, in prep.) that may
1051 help overcome the undersampling of carbonate system properties and could feed directly into
1052 DA applications.

1053 There is an important subtlety to the application of data-assimilative models when quantifying
1054 net CO₂ uptake due to OAE, which is highly relevant for MRV. When the net CO₂ uptake is
1055 quantified by calculating the difference between two simulations, one with and one without
1056 OAE (one of these is realistic, the other counterfactual), it is not appropriate to assimilate
1057 biogeochemical observations of properties affected by the alkalinity enhancement. The
1058 assimilation of alkalinity-related observations to constrain one of the simulations in the pair
1059 would eliminate the ability to make comparisons between the two. However, assimilation of
1060 observations that are unaffected by OAE (e.g., temperature, salinity, oxygen, etc.) can be
1061 applied to both simulations of the pair. Further research and method development are required
1062 to identify the best approaches for leverage DA in this context.

1063 3.4 Uncertainty analysis

1065 Model results should be paired with sound qualitative and quantitative uncertainty estimates,
1066 especially when used for practical decisions. Estimating the uncertainty of model simulations,
1067 however, is inherently difficult because typically one is most interested in simulation outputs
1068 for which observations are not available (e.g., unobserved or insufficiently observed properties
1069 or fluxes in the past, properties and fluxes in the future); hence, standard procedures and
1070 metrics for model validation (Section 3.2) are not helpful for this aspect. Uncertainty estimates
1071 could be based on extensive model parameter and configuration sensitivity studies and
1072 comparisons with models that include more realistic representations of uncertain or

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In situ measurements of the carbonate system are much more limited temporally and especially spatially than remote sensing observations from satellites, which are the backbone of physical (satellite sea surface temperature and sea surface height) and biogeochemical data assimilation (see above). Observations of pH, pCO₂, alkalinity, and DIC used to be limited to moorings and research cruises but have more recently been extended by automated observing systems, such as gliders, BGC-Argo floats and uncrewed surface vehicles (Bushinski et al. 2019).

1090 parameterized processes. Furthermore, since specification of uncertainty is an integral part of
1091 DA, DA methodologies provide a useful framework for estimating uncertainty, especially
1092 ensemble-based methods.

1093
1094 Any DA application requires uncertainty specification of the observations that are assimilated
1095 and can provide uncertainty estimates of the results of the assimilation procedure. Specification
1096 of uncertainty in the input data is necessary to inform the DA machinery about how much
1097 weight and reach each data point or data type should have in influencing the outcome. The
1098 more realistic the uncertainties of the input data, the better the DA outcomes in terms of
1099 explanatory or predictive skill. It is important to note that “better” does not mean more precise
1100 in this context. Overconfidence in the accuracy of assimilated observations will lead to
1101 overfitting and a degradation of predictive skill. In the case of parameter optimization, the
1102 output of the assimilation exercise is a set of optimized parameters. The uncertainty of optimal
1103 parameters, referred to as *a posteriori* errors, is determined by a Hessian analysis of the cost
1104 function in combination with the uncertainty of the input parameters before optimization, the
1105 so-called *a priori* errors (Thacker et al. 1989, Fennel et al. 2001). In the case of ensemble-based
1106 state estimation, the ensemble spread of the reanalyzed model state provides a spatially and
1107 temporally resolved estimate of the uncertainty of the reanalysis (Yu et al. 2018, Hu et al. 2012).

1108
1109 However, an important caveat is that subjectivity enters the uncertainty specification in all of
1110 these approaches. For example, in the case of parameter optimization the assumed *a priori*
1111 errors, their probability distributions, and the choice of the cost function are subjective and
1112 influence the *a posteriori* errors (but interestingly the values of the observations themselves do
1113 not). In the case of ensemble-based state estimation, the sources of uncertainty inherent in the
1114 model simulation have to be specified and simulated by generating variations within a model
1115 ensemble. Sources of uncertainty include errors in atmospheric forcing and boundary
1116 conditions, model parameters, and structural uncertainty. Uncertainty in forcing and boundary
1117 conditions is often represented by perturbing the time of sampling, uncertainty in parameters is
1118 represented by sampling from a probability distribution (based on *a priori* assumptions about
1119 the uncertainty of each parameter), and the structural uncertainty is typically represented via
1120 brute-force inflation factors that amplify ensemble spread. Yu et al. (2019), Li et al. (2016), and
1121 Thacker et al. (2012) provide examples where different sources of model uncertainty are
1122 accounted for. While the mechanics by which the model ensemble is generated and spreads
1123 over time is thus subjective, grossly inappropriate choices will lead to obviously wrong or
1124 degraded reanalyses. The success of a DA exercise, which is best judged by an evaluation of
1125 whether the predictive power of the model has improved, thus provides a useful reality check
1126 on whether the choices for specifying uncertainty were appropriate.

1127
1128 How can the framework for specifying and estimating uncertainty from model ensembles be
1129 applied in the context of OAE research? Two different cases should be considered here: 1)
1130 model applications where the absolute value of quantities matters for the research question to
1131 be addressed and thus the uncertainty of the simulated output, and 2) applications where
1132 information about the difference between a simulation with and without OAE is of interest and

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1134 the uncertainty of this difference (e.g., the net CO₂ uptake and its uncertainty in the context of
1135 MRV). Examples of the first case include studies of the stability of added alkalinity (i.e.,
1136 simulation of runaway calcium carbonate precipitation) and studies about the exposure of
1137 planktonic and benthic communities to high pH. In this case, the ensemble framework
1138 described above can be applied with the caveat that the specification of all the relevant sources
1139 of uncertainty is by no means trivial and subjective to some degree.

1140
1141 The second case is highly relevant for MRV of OAE where one is interested in accurately
1142 quantifying the increase in seawater DIC due to OAE with well characterized uncertainty. In
1143 this case, one would use two simulations that are based on an identical model set-up with only
1144 one difference, namely a source of alkalinity is applied to one (i.e., one of these two simulations
1145 is counterfactual or hypothetical, the other would typically be as realistic as possible). It may be
1146 tempting, and is conceptually straightforward, to apply the ensemble framework for each
1147 model of the pair and combine the resulting uncertainties via error propagation. However, in
1148 practice this would not provide meaningful estimates because there are sources of uncertainty
1149 that are unaffected by OAE (e.g., atmospheric forcing) and accounting for them may
1150 significantly overestimate uncertainty in the estimated net CO₂ uptake. A more appropriate
1151 approach would be to construct an ensemble of model pairs that explicitly accounts for
1152 uncertainty related to the impacts of alkalinity addition. How to specify and simulate the
1153 sources of uncertainty directly resulting from OAE in practice remains an open research
1154 question.

1155 1156 **4 Model experimentation**

1157 In this section, we lay out general objectives for model experimentation in the context of OAE
1158 research and provide a short historical view of how these model studies have evolved (Section
1159 4.1) followed by specific recommendations for Observing System Simulation Experiments
1160 (Section 4.2) and model intercomparisons (Section 4.3).

1161 1162 **4.1 General objectives of model experimentation**

1163 General objectives of OAE modeling include (1) gaining a better understanding of the
1164 biogeochemistry of OAE, including its effectiveness and side effects, (2) supporting
1165 experiments, field trials, or commercial deployments including through the optimization of
1166 observing systems, (3) assessing global carbon-cycle and climate feedbacks, (4) understanding
1167 the role that OAE can play in climate mitigation efforts, and (5) supporting monitoring,
1168 reporting, and verification activities. At a conceptual level, model approaches for OAE can be
1169 classified as belonging into one of two groups: idealized or realistic. Idealized modeling
1170 approaches are typically driven by research questions of a fundamental nature and aim to
1171 develop or test hypotheses or provide improved process understanding while strongly
1172 simplifying a range of potentially complicating factors. They are useful for illustrating cause-
1173 and-effect relationships and the range of plausible outcomes given strong assumptions. In
1174 contrast, realistic modeling approaches aim to include a broad range of contributing factors as
1175 accurately as possible and provide detailed hindcasts or predictions that, if the model has skill,

1176 can be used for a range of practical applications. In practice, the dividing line between idealized
1177 and realistic models is blurry. Of course, no model will ever simulate all aspects of reality,
1178 hence even realistic simulations make many assumptions and are prone to errors from multiple
1179 sources. It can be effective to apply idealized and realistic approaches in a complementary
1180 manner and iteratively.

1181 It is illustrative to review briefly how modeling for OAE research has developed over the course
1182 of the last decade. Much of the early work on OAE used idealized models. Model simulations
1183 were designed to investigate whether the theoretical concept of OAE could remove large
1184 amounts of CO₂ on the global scale. Rather than trying to account for the technical and socio-
1185 economic constraints of OAE deployment, the model experiments were designed to investigate
1186 what would happen if surface alkalinity was homogeneously increased by massive amounts via
1187 a constant addition rate over extremely large regions of the ocean, e.g., in all sea-ice free waters
1188 (Paquay and Zeebe, 2013; Keller et al., 2014; Ilyina et al., 2013; Köhler et al., 2010; Köhler et al.,
1189 2013). These simulated OAE deployments will never be realized, but the model results
1190 suggested that OAE can be viable as a CDR approach. A particular advantage of this idealized
1191 approach is that the effect of OAE was easy to detect against internal model variability, i.e., the
1192 signal to noise ratio is high. The next steps in modeling OAE have remained idealized but have
1193 begun to introduce more constraints and better mechanistic or empirically derived components
1194 as experimental OAE data becomes available. Recently, modeling studies tailored to specific
1195 regions and modes of application have been conducted to support field trials or commercial
1196 deployment (Mongin et al. 2021, Wang et al. 2023). These applications must be as realistic as
1197 possible. None of the modeling studies published to date have simulated an actual OAE field
1198 trial.

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1199 4.2 Recommendations for Observing System Simulation Experiments (OSSEs)

1200 Observing system simulation experiments (OSSEs) use data-assimilative simulations to design
1201 new, or modify existing, observing systems such that deployments of observing assets, e.g.,
1202 floats, gliders, moorings, or surface vehicles, is optimized. General overviews and best practices
1203 for OSSEs are provided by Halliwell et al. (2015) and Hoffman and Atlas (2016). Examples of
1204 applications to biogeochemical models include Ford (2021), Wang et al. (2020), and Denvil-
1205 Sommer et al. (2021). Their goal is to maximize the information gained from a new or modified
1206 observing system, while keeping the number of required instruments, sensors, or deployments
1207 – and thereby cost and effort – low. OSSEs are especially valuable tools in the context of OAE
1208 research because the marine carbonate system is still undersampled, observing systems need to
1209 be designed and expanded, and new instruments deployed and configured (Boyd et al. 2023).

1210 In practice, this is done with the help of a pair of two different models or model versions, also
1211 referred to as twin experiments, as follows. A simulation of one of the models is considered to
1212 be the “truth.” This simulation is also referred to as the “nature run” and synthetic observations
1213 are generated by subsampling this nature run. This subsampling can be repeated with different
1214 sampling schemes (e.g., different variable types, different numbers of profiles, transects, and/or
1215 fixed location time series, etc.) to represent different configurations of the observing system.

1217 Finally, the synthetic observations are assimilated into the other model for which a non-
1218 assimilative simulation, the so-called “free run,” is also available. The skill of this data-
1219 assimilative simulation, also referred to as the “forecast run,” can be assessed against the free
1220 run using independent observations that are also sampled from the nature run. In this way the
1221 impact of different sets of observations on the data-assimilative model can be measured and
1222 assessed.

1223 While conceptually straightforward, care and consideration are required when setting up
1224 OSSEs. For example, the choice of the two model versions making up the twin is important. If
1225 the models chosen for the truth and forecast runs are versions of the same model
1226 implementation that were generated by perturbing initial, forcing or boundary conditions in
1227 one of them, the method is referred to as the “identical twin” approach. If two different model
1228 types are used, they are “non-identical twins.” The intermediate approach where the same
1229 model type is used but in different configurations (e.g., different physical parameterizations
1230 and/or spatial resolution) is referred to as fraternal twin. The identical twin approach has been
1231 more common in oceanic DA applications although atmospheric OSSEs have shown that it can
1232 provide biased impact assessments (Hoffman and Atlas, 2016) typically because the error
1233 growth rate between the truth and forecast runs is insufficient. A direct comparison of the non-
1234 identical and identical twin approach for an ocean circulation model of the Gulf of Mexico has
1235 been conducted by Yu et al. (2019). In their assessment of the impacts of the existing observing
1236 system (consisting of satellites and Argo floats), the identical twin approach provided overly
1237 optimistic improvements in model skill after assimilation of data from some observing assets
1238 (specifically sea-surface height and temperature) but undervalued the contribution from
1239 temperature and salinity profiles. They concluded that skill assessments and OSSEs using the
1240 non-identical twin approach are more robust. Similar concerns likely apply to OSSEs for
1241 biogeochemical properties, but this remains to be studied systematically.

1242 **4.3 Recommendations for intercomparisons**

1243 A common approach to assessing model uncertainty are coordinated, multi-model studies,
1244 commonly called model intercomparison projects or MIPs. They can be used to explore the
1245 simulated range of model behaviors, to isolate the strengths and weaknesses of different models
1246 in a controlled setting, and to interpret, through idealized experiments, inter-model differences
1247 (IPCC 2013). Carefully designed experiments can also offer a way to distinguish between errors
1248 particular to an individual model and those that might be more universal and should become
1249 priority targets for model improvement (IPCC 2013). These studies rely on common agreed-
1250 upon protocols for simulating certain processes and writing of diagnostic output to ensure that
1251 best practices are followed, and results are comparable (e.g., Griffies et al., 2016). The best-
1252 known model intercomparison project is probably the Coupled Model Intercomparison Project
1253 (CMIP, Eyring et al., 2016), which is currently finishing up its 6th phase. Within CMIP6, the
1254 carbon dioxide removal intercomparison project (CDRMIP; Keller et al., 2018) is the first project
1255 to develop a model intercomparison experiment for ocean alkalinity enhancement. This and
1256 other MIP examples, including those conducted at smaller region scales (Wilcox et al., 2022),
1257 provide a blueprint for developing coordinated multi-model experiments.

1258 The following key practices have proven useful in previous coordinated multi-model
1259 comparisons. Since broad participation is typically desired, the protocol should be
1260 straightforward for modeling groups to implement, otherwise few will have the resources to
1261 participate. In practice this means avoiding new implementations of complex code or requiring
1262 too many or too long simulations. If applicable, forcing data should be centrally prepared and
1263 provided to participants in a standardized way that enables easy modification or reformatting,
1264 if needed, for use with different models. Using common simulations that modeling groups are
1265 likely to have completed already, e.g., climate change scenarios, as control runs and
1266 experimental branching points is helpful for minimizing the number of additional required
1267 simulations. It is useful to establish common practices that facilitate the production and analysis
1268 of the model output, e.g., what should be archived and shared (Juckes et al., 2020) and data
1269 standards governing the structure and required metadata for model output (Pascoe et al., 2020).
1270 Shared software to standardize model output, such as the Climate Model Output Rewriter
1271 (CMOR; <https://cmor.llnl.gov/>) commonly used in CMIP, can be helpful. To maximize the use of
1272 model output, it should be made available for public download with digital object identifiers
1273 (DOIs). The Earth System Grid Federation (ESGF) is an example of such a system (Petrie et al.,
1274 2021). If applicable, preparing and providing quality-controlled observational datasets for
1275 model evaluation is useful for facilitating analytical efforts (Waliser et al., 2020). Coordinating
1276 the analysis is helpful to avoid duplicative efforts and ensure consistent application of
1277 evaluation metrics. Finally, the design of a coordinated multi-model experiment and all its
1278 procedures should be well documented in publications or permanently archived protocols. It
1279 is advisable to test the multi-model experiment with a small subset of models, before inviting a
1280 large number of participants. Furthermore, it is worth remembering that the science questions
1281 must be appropriate. MIPs require much effort and not every science question needs a MIP to
1282 be answered.

1283 5 Summary and Key Recommendations

1284 A range of modeling tools and analysis methods are available for OAE research to address
1285 questions from micro- to global scales; however, each of these tools and methods has limitations
1286 and caveats that model users and users of model-generated outputs need to be aware of.
1287 Furthermore, this new field of research poses questions and challenges that current tools were
1288 not designed to address, necessitating further development.

1289
1290 A common objective of all modeling approaches described in this [article](#) is to simulate the
1291 spatio-temporal evolution of carbon chemistry properties in seawater by accounting for the
1292 physical, chemical, and biological processes that determine this evolution. Idealized models,
1293 which neglect some aspects of reality in the interest of simplicity and clarity of assumptions,
1294 have long been used to test basic questions about OAE. As research questions are becoming
1295 more focussed on the practical aspects, feasibility, and ecosystem impacts of OAE, more realistic
1296 models are increasingly desirable. A skillful realistic model can provide spatial and temporal
1297 context for observations, including estimates of properties and fluxes not directly observed.
1298 Such model will include parameterizations of the relevant processes for the research objective to
1299 be addressed and will be constrained by observations that contain sufficient meaningful

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1314 information. However, model formulations of several properties and processes relevant to OAE
1315 research remain uncertain or highly simplified. For example, presently used model
1316 representations of alkalinity in seawater are likely inadequate and may require explicit
1317 representation of at least some of the multiple biotic and abiotic sources and sinks of alkalinity;
1318 the mechanisms and triggers for spontaneous calcium carbonate precipitation are only
1319 beginning to be described and not yet represented in models; and the impacts of pH
1320 perturbations on plankton diversity and trophic interactions remain an active area of study and
1321 unaccounted in biogeochemical models. Furthermore, it is difficult to obtain solid constraints on
1322 the seawater carbonate system, especially in sufficient spatial and temporal resolution for
1323 robust model validation and DA. Theoretically, knowledge of two of the carbonate system
1324 parameters allows calculation of the others, but unfortunately $p\text{CO}_2$ and pH, the pair most
1325 accessible to autonomous measurement, results in high uncertainties.

1326 One inherent challenge to OAE research is the multiscale nature of many of the relevant
1327 questions. Different modelling tools are available for different spatial scales. While some
1328 research questions may fall neatly within the limited spatial range of a particular model, many
1329 do not and require a bridging of scales that could be accomplished via new parameterizations
1330 yet to be developed or dynamic coupling of different modeling tools. It is important to
1331 emphasize that models have to be tailored to the questions they are meant to address. This
1332 means considering what level of model complexity is required and seeking parsimonious
1333 representations that are well-supported by empirical constraints.

1334
1335 It is important to note that even after thorough validation, any model simulation will be prone
1336 to errors and uncertainties due to inaccuracies in model inputs, structural uncertainty due to
1337 numerical approximation schemes and insufficient process understanding or representation,
1338 and inaccurate model parameters and parameterizations. Deviations between models and
1339 reality can be reduced by DA, which is typically applied either to systematically optimize
1340 model parameters or to produce optimal estimates of the ocean state. Optimization of model
1341 parameters addresses systematic model errors and biases; it is useful for systematic testing of
1342 different model formulations during model design. State estimation assumes an unbiased
1343 model and addresses unresolved stochasticity, thus leading to model states that are in better
1344 agreement with the observed ocean state. However, successful application of DA critically
1345 requires sufficient observations. Currently, the biggest impediment to implementing data-
1346 assimilative models for OAE research is the sparsity of carbonate system observations. OSSEs,
1347 data-assimilative simulations that inform how to place observing assets most effectively, will
1348 prove useful in this context. It should also be noted that assimilation of carbonate system
1349 parameters is not appropriate when models are applied for MRV.

1350
1351 Uncertainty analysis is a necessary component of any quantitative research and will be an
1352 essential deliverable for effective approaches to MRV. Ensemble-based DA methodologies
1353 provide a useful framework for estimating uncertainty. Consideration of this framework
1354 illustrates the “law of conservation of difficulty” applies here. Quantitative assumptions about
1355 the uncertainty distributions of input data and input parameters, and of structural uncertainties

1356 inherent in the model are required to obtain an uncertainty estimate of the model output, in
1357 other words, difficult assumptions about errors have to be made somewhere. A common
1358 approach to assessing model uncertainty is by coordinated, multi-model intercomparison. Such
1359 studies can be used to explore the range of simulated behaviors and the strengths and
1360 weaknesses of different models and, by elucidating inter-model differences, they can offer
1361 guidance on priority targets for model improvement.

1362

1363 Key recommendations arising from this article are as follows:

1364

1365 • Idealized models of particle-fluid interaction are recommended to address questions
1366 about dissolution and precipitation kinetics at the scale of particles, realistic local-scale
1367 models are recommended for addressing questions about nearfield processes in the
1368 turbulent environment around injection sites, and larger-scale regional or global ocean
1369 models are recommended to support observational design for field experiments, to
1370 demonstrate possible verification frameworks, and to address questions about global-
1371 scale feedbacks on ocean biogeochemistry.

1372 • When simulating OAE approaches that may generate high oversaturation with respect
1373 to carbonate, spontaneous precipitation of carbonates needs to be considered and
1374 appropriate approaches should be developed, e.g., using near-field models to
1375 mechanistically represent this process and a meta-model approach to develop
1376 parameterizations that are suitable for far-field and larger-scale models.

1377 • Shortcomings in current-generation models in terms of representing physiological
1378 responses of the plankton community to OAE (especially when using crushed-rock
1379 feedstocks) need to be recognized, better qualified, and addressed. Empirical research
1380 exploring physiological sensitivities should be used to develop prioritizations of key
1381 model processes comprising early targets for implementation.

1382 • The exchange of solutes between the sediments and overlying water influences the
1383 seawater carbonate system with DIC from the remineralization of organic matter being
1384 returned to overlying water (and alkalinity if this remineralization occurs anaerobically),
1385 dissolution of CaCO₃ releasing alkalinity, and burial of CaCO₃ acting as alkalinity sink.
1386 Accounting for these exchanges between sediments and overlying water and its
1387 variability on tidal, seasonal, interannual, and millennial timescales will likely be
1388 necessary in regional and global biogeochemical models that aim to simulate alkalinity
1389 cycling.

1390 • River inputs of alkalinity and DIC in regional and global ocean biogeochemical models,
1391 including fluxes resulting from land-based CDR applications, should be accurately
1392 accounted for. Efforts should be made to improve quantification of riverine fluxes
1393 resulting from ongoing field trials and commercial applications, and to establish
1394 initiatives to effectively track the solute additions from terrestrial alkalinity
1395 enhancements.

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- When simulating large-scale deployment of OAE in ocean-only models with prescribed atmospheric CO₂, the subtle changes in the atmospheric CO₂ inventory resulting from CDR should be accounted for.
 - Models should be tailored to the specific questions they are meant to address while seeking parsimonious representations that are well-supported by empirical constraints. For example, for the purpose of MRV it may be appropriate to neglect biological dynamics since the core target is to capture the net air-sea CO₂ exchange associated with the OAE-induced surface ocean pCO₂ anomaly.
 - Model validation should be an integral part of model implementation and application. For OAE research, validation is a two-step challenge. First, it is necessary to validate unperturbed model baselines to gain confidence that the natural variability is represented appropriately and to quantify model uncertainties. Second, it should be verified that the model accurately represents the perturbations of an alkalinity addition.
 - Since no single model validation metric provides a complete picture of a model's skill, multiple complementary metrics should be used in combination. Furthermore, different points in space and time, and a breadth of variable types should be part of any comprehensive validation.
 - Data assimilation, the process of improving the dynamical behavior of models by statistically combining them with observations, should be employed in order to obtain the most accurate model simulations possible, e.g., to optimize model parameters or to estimate the ocean state. The former addresses systematic errors and biases in models, while the latter assumes an unbiased model and addresses unresolved stochasticity.
 - When applying data-assimilative models for quantification of the OAE-induced net CO₂ uptake by calculating the difference between a realistic and a counterfactual simulation, it is not appropriate to assimilate biogeochemical observations of properties affected by the alkalinity enhancement as this would eliminate the ability to make valid comparisons between the two simulation. However, assimilation of observations that are unaffected by OAE can be applied to both simulations of the pair.
 - Successful application of DA critically requires sufficient observations either of the properties that the model parameters to be estimated depend on or of the state variables that are being estimated. Observing System Simulation Experiments are recommended to design observing strategies tailored to the needs of specific OAE applications.
 - Model results should be paired with sound qualitative and quantitative uncertainty estimates, especially when used for practical decisions. DA methodologies provide a useful framework for estimating uncertainty, especially ensemble-based methods. Another common approach to assessing model uncertainty are coordinated, multi-model studies, commonly called model intercomparison projects or MIPs.

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